

**St. Lawrence River Sediment
Chemical Assessment 1997,
Cornwall, Ontario**

September 1999



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**Ministry of the
Environment**

St. Lawrence River Sediment Chemical Assessment 1997, Cornwall, Ontario

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FOREWORD

Since 1970 there have been several sediment surveys in the St. Lawrence River along the Cornwall waterfront to delineate the extent of sediment contamination and to associate these contaminants with local point sources. These surveys identified mercury contaminated sediment extending downstream from Domtar Fine Paper Ltd. and ICI Forest Products (formerly called CIL), and adjacent to, and downstream of the former Courtaulds Fibres Canada property (herein referred to as Courtaulds) (MOE 1979; Kauss et al. 1988; Anderson 1990; Richman 1994; Richman 1996). Sediment samples collected in both these areas since 1970 exceeded the "severe effect level" (SEL) of the Ontario Sediment Quality Guidelines for mercury and sediment collected downstream of Courtaulds also exceeded the SEL for lead, copper and zinc.

This project was put forward by the Cornwall Sediment Management Plan committee to provide information for the development of a plan to address contaminated sediment in the St. Lawrence River at Cornwall, Ontario. The data from this report will be integrated with additional information generated by related studies in the St. Lawrence River so that sediment management options can be proposed.

This report has been prepared under the auspices of the Canada-Ontario Great Lakes Remedial Action Plan Program. Financial support for the sampling projects, data analysis and report writing was provided by Environment Canada (EC) and the Ontario Ministry of Environment (MOE). This was a collaborative project between MOE (Environmental Monitoring and Reporting Branch and Eastern Region) and EC (Environmental Conservation Branch; Restoration Program Division and National Water Research Institute).

The report presents the findings, recommendations, and conclusions of the author, and does not necessarily represent the views or policies of the supporting agencies.

For additional technical reports or information on the St. Lawrence River Remedial Action Plan (RAP), contact the Ontario Ministry of Environment at 133 Dalton Road, Kingston Ontario. K7L 4X6.

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Hans Biberhofer (EC) is acknowledged for the report figures and his assistance with survey design, participation in the field and coordination of sample delivery to EC laboratories for analysis.

The author would also like to thank Jeff Ridal of the St. Lawrence River Institute of Environmental Studies for providing access to their laboratory for sample preparation and storage.

Laboratory sample analyses were performed by EC National Laboratory for Environmental Testing (NLET) and MOE Laboratory Services Branch.

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SUMMARY

The St. Lawrence River near Cornwall, Ontario was designated as a Great Lakes Area of Concern (AOC) in 1985 by the International Joint Commission in part because of contaminated sediment located along the north shore of the Cornwall waterfront. Sediment surveys along the waterfront since 1970 showed sediment concentrations greater than the "severe effect level" (SEL) of the Ontario Sediment Quality Guidelines for mercury extending downstream from Domtar Fine Paper Ltd. (herein referred to as Domtar) and ICI Forest Products (formerly called CIL). Sediment downstream of the former Courtaulds Fibres Canada facility (herein referred to as Courtaulds) had concentrations greater than the SEL for mercury, lead, copper and zinc (MOE, 1979; Kauss et al. 1988; Anderson 1990; Richman 1994; Richman 1996). As such, a Remedial Action Plan (RAP) to improve the local conditions of the aquatic environment identified the need to address the issue of sediment contamination along the Cornwall waterfront (Dreier et al. 1997).

Prior to the development of a sediment management strategy for the waterfront more recent information was required on the local sediment quality to identify areas of contaminated sediment which may require consideration for remediation.

The objectives of the 1997 sediment survey were as follows:

- (1) *to determine if concentrations of various metals and organics in sediment located in a deposition zone about 1.4 km downstream of the Domtar/ICI diffuser and adjacent to the north east side of Cornwall Island exceed the Provincial Sediment Quality Guidelines "severe effect level" (SEL) and "lowest effect level" (LEL).*
- (2) *to update sediment quality information downstream of the Courtaulds facility using sampling sites based on a 1994 MOE sediment survey.*

Sediment was collected from 24 stations using a mini-box corer (Figure 1). At each station 10 cm cores (using 10 cm core tubes, diameter-6.5 cm) and surface samples (top 3 cm) were collected from the mini-box corer. Sediment samples were analysed for trace metals, phosphorus, % TOC and particle size.

All sediment samples collected along the Cornwall waterfront (north shore of the channel) exceeded the LEL for mercury and 46 % of the samples exceeded the SEL. The highest concentrations of mercury in 10 cm core samples (19.5 $\mu\text{g/g}$ and 11.2 $\mu\text{g/g}$) were detected in sediment collected downstream of Courtaulds at stations CS131 and CS128 respectively. In contrast, all stations on the south side of the channel had mercury concentrations less than the LEL indicating that this area was not contaminated. These results were consistent with previous sediment surveys (MOE 1979; Kauss et al. 1988; Anderson 1991; Richman 1994, 1996).

A comparison of metal sediment concentrations in samples (10 cm cores) collected from the north side of the north channel showed that median concentrations of cadmium (0.89 $\mu\text{g/g}$), nickel (24 $\mu\text{g/g}$) and TOC (2.5%) were similar to concentrations detected on the south side of the channel (1.02 $\mu\text{g/g}$, 30.8 $\mu\text{g/g}$ and 2.5% respectively). Median concentrations of copper and lead were also similar on both sides of the channel, however, the range in concentrations for copper and lead was higher on the north shore than the south. Stations with the highest concentrations of copper and lead were located downstream of Courtaulds. Median concentrations of zinc (186 $\mu\text{g/g}$) were higher on the north shore compared with the south shore (136 $\mu\text{g/g}$) indicating enrichment for this parameter as would be expected based on historical discharge patterns from Courtaulds and confirmed in previous sediment surveys (Kauss et al. 1988; Anderson 1991; Richman 1994, 1996).

A comparison of particle size corrected data from Cornwall with data from an upstream reference area showed that the sediment along the north side of the channel was enriched with zinc at all stations and with lead and copper at selected stations downstream of Courtaulds which historically discharged these metals (MOE 1992a). Particle size corrected data showed that all stations along the north shore (downstream of the Domtar/ICI diffusers and downstream of Courtaulds) were enriched with mercury compared with the upstream reference area. All three industries were known dischargers of mercury (MOE 1979; MOE 1992a&b). Particle size corrected data also showed the same pattern described above (i.e. higher concentrations on the north shore), when comparing concentrations of mercury, lead, zinc and copper in sediment from the north side relative to sediment concentrations on the south side of the channel, confirming that the south side of the channel was not impacted by local industrial sources.

A statistical examination of the relationships among contaminant concentrations at all sampling locations (Principal Component Analysis), grouped Hg, Pb, Zn and Cu together indicating that these parameters were correlated with one another (i.e. exhibited similar patterns of variation from one location to another). Stations with high Hg, Pb, Zn and Cu sediment concentrations were also grouped together and were all located downstream of Courtaulds which historically discharged these metals.

An assessment of contaminant concentrations in the bottom 10 cm of selected cores confirmed that stations about 1.4 km downstream of the Domtar/ICI diffusers were historically contaminated by mercury. These results were consistent with data from the 1970's (MOE 1979; Kauss et al. 1988). The high concentrations (4.63 $\mu\text{g/g}$ and 13.70 $\mu\text{g/g}$) were likely due to upstream discharges of mercury from the Fly Creek sewer and from the Domtar/ICI diffuser when it came on line in 1972 and replaced the Fly Creek sewer. Surface samples from these stations show some enrichment of mercury relative to the south side of the north channel (which was not impacted by local sources) and relative to upstream reference stations, however, present concentrations were below the SEL and lower than contaminated sites downstream of Courtaulds.

An analysis of covariance (ANCOVA) was used to compare metal concentrations in the 1997 surface sediment samples (top 3 cm) with the concentrations in the 1997 core samples (top 10 cm) collected from the same stations to review changes in contaminant concentrations over time. This was based on the assumption that the top 3 cm of sediment represented the most recent sediment quality. The ANCOVA found that there was no significant difference in metal concentrations in the surface samples compared with the 10 cm core top samples for all the parameters tested with the exception of cadmium and chromium which showed weak statistical differences. This suggests that sediment quality has not changed at the surface relative to the underlying sediment. However, sediment mixing due to bioturbation and other physical processes within the river may make any change in sediment quality too small to measure within the top 10 cm.

CONCLUSIONS

- 1) *Surface sediment located in the deposition zone about 1.4 km downstream of the Domtar/ICI diffuser has mercury concentrations that were greater than the LEL but less than the SEL. Sediment at this site was historically contaminated with mercury. This is evident from high concentrations in sediment collected from the bottom of the core (4.63 ug/g and 13.70 ug/g) relative to mercury concentrations in sediment at the top of the core sample (1.19 ug/g and 1.71 ug/g respectively) . This area was not contaminated with other metals.*
- 2) *Sediment located adjacent to the north east side of Cornwall Island was not contaminated with metals.*
- 3) *The highest concentrations of mercury, lead, copper and zinc in the study area were located downstream of the former Courtaulds Fibre Canada facility. Remediation of contaminated sediment, if considered, should be focussed in this deposition zone. Sampling sites in this area could be used for future monitoring to assess changes in sediment quality with time.*

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INTRODUCTION

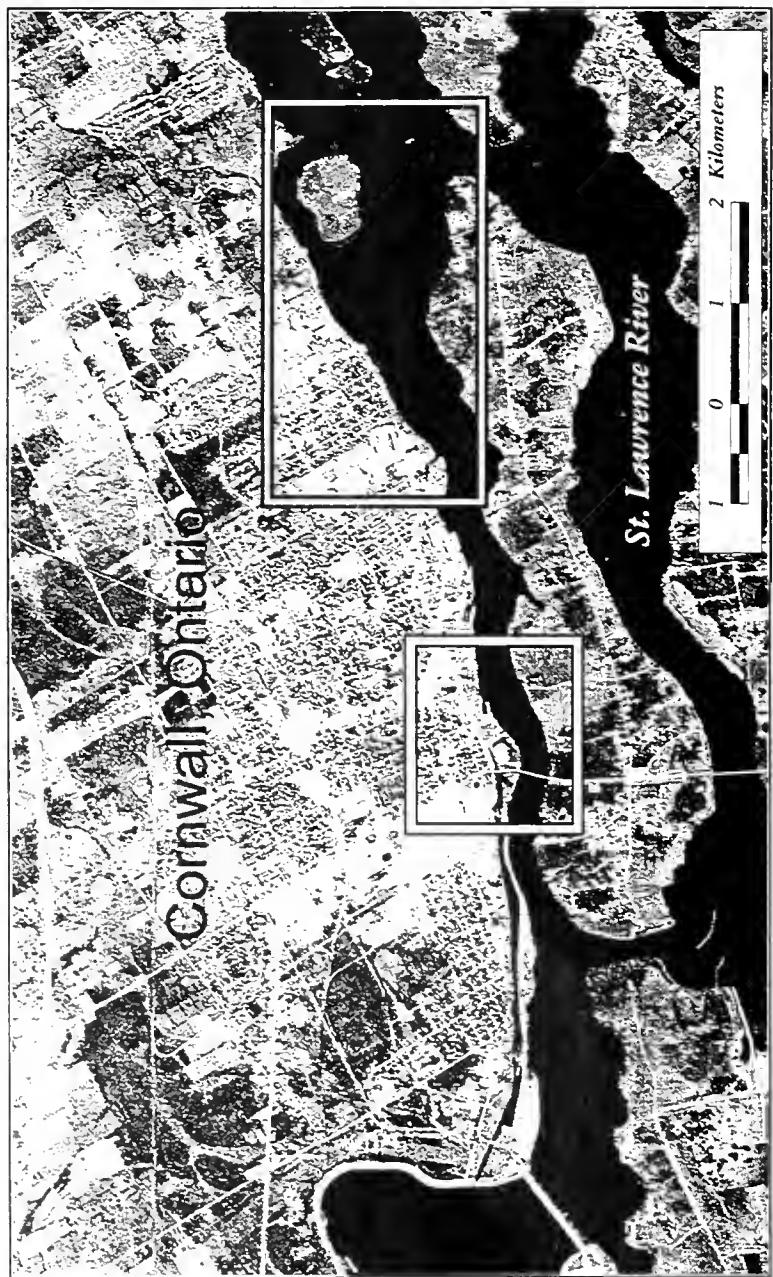
Since 1970 there have been several sediment surveys in the St. Lawrence River along the Cornwall waterfront to delineate the extent of sediment contamination and to associate these contaminants with local point sources. These surveys identified mercury contaminated sediment extending downstream from Domtar Fine Paper Ltd. (herein referred to as Domtar) and ICI Forest Products (formerly called CIL), and adjacent to, and downstream of the former Courtaulds Fibres Canada property (herein referred to as Courtaulds) (MOE 1979; Kauss et al. 1988; Anderson 1990; Richman 1994; Richman 1996) (Figure 1). Sediment samples collected in both these areas since 1970 exceeded the "severe effect level" (SEL) of the Ontario Sediment Quality Guidelines for mercury and sediment collected downstream of Courtaulds also exceeded the SEL for lead, copper and zinc.

Local industries located upstream of the areas with contaminated sediment have discharged mercury, zinc, lead and copper to the St. Lawrence River. Based on a six month average loading calculation from October 1989 to March 1990, Courtaulds Fibres Canada was one of the largest dischargers of zinc in Ontario. Courtaulds discharged 399.95 kg/d of zinc, 2.25 kg/d of lead, 0.726 kg/d of copper and 0.120 kg/d of mercury (MOE 1992a-unpublished data). Historically Courtaulds discharged higher concentrations of lead because of the corrosion of their lead pipes and their use of lead as a lining in their acid storage tanks. Prior to its closure, the company had almost completely replaced its lead pipes and tank lining with plastic products reducing the concentrations of lead that were discharged to the river, however, the company still discharged higher lead concentrations than the other local upstream sources in the study area (MOE 1992a,b). Courtaulds also decreased its mercury loadings to the river due to changes in their distributors of caustic soda and sulphuric acid which were contaminated with mercury, however, concentrations in their effluent were still higher than other local sources when data was collected in 1989. In November 1992, Courtaulds Fibres Canada closed its facility in Cornwall, Ontario.

Domtar discontinued its use of mercurial slimicides in 1964 (MOE 1979) and according to six month average loading calculation from January 1990 to June 1990, Domtar discharged 5.7 kg/d of zinc and 0.843 kg/d of copper (MOE 1991). ICI significantly reduced its loadings of mercury to water in the early 1970s to an average of 23 kg/yr (0.063 kg/d) from a high of 590 kg/yr (MOE 1979; MOE 1992b). ICI discharged insignificant amount of zinc (0.111 kg/d) relative to Courtaulds and 0.178 kg/d of lead and 0.082 kg/d of copper. ICI closed its Cornwall chlor-alkali plant in 1995.

The St. Lawrence River near Cornwall, Ontario was designated as a Great Lakes Area of Concern (AOC) in 1985 by the International Joint Commission in part because of the mercury contaminated sediment. As such, a Remedial Action Plan (RAP) to improve the local conditions

Figure 1: Study area
and sampling sites.



of the aquatic environment identified the need to address the issue of sediment contamination along the Cornwall waterfront (Dreier et al. 1997).

Prior to the development of a sediment management strategy for the waterfront more recent information was required on the sediment quality located in a deposition zone about 1.4 km downstream of Domtar and ICI. As well, this survey provided an opportunity to reassess the sediment quality of a deposition zone on the south side of the north channel adjacent to Cornwall Island identified through the use of the RoxAnn™ seabed classification system (Rukavina 1997). This area was sampled previously by the Ministry of Environment (MOE) and was shown not to be contaminated by local point source discharges (MOE, 1979; Kauss et al. 1988; Anderson 1990).

The objectives of the 1997 sediment survey were as follows:

- (1) *to determine if concentrations of various metals and trace organics in sediment located in a deposition zone about 1.4 km downstream of the Domtar/ICI diffuser and adjacent to the north east side of Cornwall Island exceed the Provincial Sediment Quality Guidelines "severe effect level" (SEL) and "lowest effect level" (LEL).*
- (2) *to update sediment quality information downstream of the former Courtaulds Fibres Canada facility using sampling sites based on a 1994 MOE sediment survey.*

Accordingly, sediment was collected jointly by MOE and Environment Canada (EC) from 24 stations in October 1997, and samples were analysed for a range of contaminants and physical parameters.

METHODOLOGY

Sampling Stations and Field Methods

Sediment was collected from 24 stations using a mini-box corer from October 21 to 23, 1997 (Figure 1). Results from detailed acoustic mapping of bottom sediment type, grain-size types and sediment thickness using a RoxAnn™ seabed classification system were used to choose sampling stations (Rukavina 1997). Appendix A provides details on exact station locations. A boat equipped with RoxAnn™ was used to identify areas with sufficient soft sediment for efficient use of the mini-box corer. These sites were marked with floats and the exact northing and easting were recorded using a differential Global Positioning System (DGPS). The sampling vessel then returned to these locations for sediment sample collection. An attempt was made to return to sites previously sampled in 1994 but deviations in the location of some of the stations from the 1994 survey grid were required to meet the needs of the mini-box corer.

At each station two 10 cm cores (using 10 cm core tubes, diameter-6.5 cm) were collected from the mini-box corer (38 x 38 x 46 cm), homogenized in a hexane rinsed glass tray using a stainless steel hexane rinsed spatula and split into three pre-cleaned 500 mL amber sample jars with Teflon lined lids. Care was taken to rinse all mixing tools and tray with water and hexane between sampling stations to avoid cross contamination. Sediment in one jar was analysed for organics, one jar for trace metals, TOC and particle size analysis, and one jar for mercury analysis. Three surface samples (top 3 cm) were collected from each station by removing sediment from the mini-box corer using a 3 cm square surface sampler. These were homogenized and placed in one jar for metals (including mercury), TOC and particle size analysis. All samples were frozen at -20°C until samples analysis.

At six stations (three stations about 1.4 km downstream of Domtar/ICI and three stations located in the deposit on the south side of the north channel), two full length cores were collected from the box corer (maximum depth of box-corer: 46 cm) to assess contamination at the bottom of the cores as well as within the top 10 cm of the cores. At these stations the top 10 cm from each full length core was sectioned and homogenized (rather than using the 10 cm core tubes) and the bottom 10 cm of the two cores were sectioned and then homogenized. The samples were split into sample jars as described above. The total lengths of these cores are provided in Appendix A.

QA/QC samples were collected at five stations to evaluate sample variability within the mini-box corer. First routine samples were collected at these stations as described above. Then a second mini-box core sample was collected and six 10 cm cores were removed. Pairs of 10 cm cores were homogenized per box corer sample resulting in a total of three replicate samples for "within box" variability analysis. This procedure was performed at four of the five QA/QC stations. The samples from three of the stations were analysed for the parameters described in Table 1, and the samples from the fourth station were archived. The fifth QA/QC station was used to evaluate laboratory analytical precision and variability due to sample handling and collection. Two of the three replicate samples collected at this station were each split three ways. The three sub-samples from each replicate were analysed for metals (excluding mercury), TOC and particle size. The third single replicate was analysed for all the parameters listed in Table 1.

In order to assess temporal trends between 1997 and 1994 data, an interlaboratory comparison of metals and particle size analysis was required since the EC lab was responsible for the 1997 analysis and the MOE lab analysed the 1994 sediment samples. At nine stations sediment samples were split for interlaboratory comparisons between MOE and EC. Two 10 cm cores were homogenized and split into two samples. Results were available for eight stations due to breakage of one sample container in transit.

Contaminant Analysis of Sediment

Sediment samples were analysed for particle size, a complete ICP metal scan, total phosphorus, % TOC and polycyclic aromatic hydrocarbons (PAHs). For selected stations, polychlorinated biphenyls (PCBs) and organochlorine pesticides were also analysed. Parameters discussed in this

report include only trace metals, nutrients and particle size. A complete list is provided in Table 1. Sediment analyses for PCBs and PAHs were not completed when this report was written.

| Table 1: Sediment samples were analysed for the following parameters: | |
|---|--|
| Total Organic Carbon (TOC) Total Phosphorus (TP) | |
| Chromium Mercury Lead Aluminum Copper | |
| Cadmium Manganese Zinc Nickel Iron | |
| Particle Size | |

Sediment samples were analysed by EC at the National Laboratory for Environmental Testing (NLET). Freeze dried sediment screened at 2 mm to remove bulk debris was weighed to 0.5 grams in a PFA Teflon vessel; 9 mL of nitric acid, 2 mL of hydrochloric acid and 1 mL of hydrogen peroxide were added. The vessel was sealed, placed in a high pressure microwave oven and the sediment was allowed to digest at 200 °C for at least 15 minutes. The digests were brought to a final volume of 100 mL with deionized water and stored in FEP Teflon bottles. The procedure was based on EPA Method 3052 which describes the digestion of siliceous and organically based matrices. This study was not concerned with the extraction of silica-bound metals, hence hydrofluoric acid was not added to the extraction mixture. Al, Ba, Cr, Cu, Fe, Mn, P, Pb, Sr, Ti, V, Zn, Ca and Mg were analysed by inductively coupled plasma-optical emission spectrometry (ICP-OES). After a ten-fold dilution of the digest, lower concentration elements, Be, Cd, Co, Li, Mo, Ni, Rb, Tl, La and U were measured by inductively coupled plasma-mass spectrometry (ICP-MS). One gram of wet sediment was digested under the same microwave conditions, and Hg was analysed by cold vapour atomic absorption spectrometry (CVAAS). The Hg results were reported on a dry weight basis after the determination of the moisture content.

Particle size was analysed by the Sieve and Sedigraph method. Sand and coarse fractions were determined by using electromagnetic 8 inch brass sieves while size fractions less than 63 µm were measured using the 5100 x-ray analyser. The two data sets were merged to provide the total particle size analysis.

For split samples analysed at the MOE Rexdale laboratory all laboratory analytical procedures for contaminants in sediment followed the methodology outlined in the Handbook of Analytical Methods for Environmental Samples (MOE 1983). Procedural updates for metals analysis are

provided in MOE 1989a & b. Procedural updates for nutrient analysis and TOC are provided in MOEE 1995a & b.

Data Analysis

Identification of Contaminated Sediment

Sediment contaminant concentrations were compared with the Provincial Sediment Quality Guidelines (Persaud et al. 1992). These guidelines describe three "effect" levels for different contaminants in terms of potential effects on the benthic community: (1) the no observed effect level; (2) the lowest effect level (LEL) which is the level of sediment contamination that can be tolerated by the majority of benthic organisms (concentrations greater than this level indicate that the benthic communities in these areas may be impaired); and (3) the severe effect level (SEL) which is the sediment concentration of a compound that is expected to be detrimental to the majority of benthic species. Sediment contaminant concentrations may exceed the LEL and/or the SEL with no apparent impact on the benthic community, nevertheless these guidelines serve as a point of reference to investigate the extent of sediment contamination within the study area and to compare the relative contamination with other locations.

To compare sediment contaminant concentrations between the north and south side of the channel the median and range in contaminant concentrations were presented.

A principal components analysis using the Numerical Taxonomy and Multivariate Analysis System (PCA) (NTSYS-pc; Rohlf 1988) was used to reduce the data onto three components representing the variables that best correlate with one another and explain the major sources of variation in sediment quality among the 24 stations. The PCA included the 3 cm surface samples and surface core samples (10 cm) from all stations. All data were log transformed with the exception of the particle size data. The PCA was repeated using non-transformed data, using only north shore data and with only the core data. The overall conclusions interpreted from the analysis remained the same for all scenarios.

Particle Size Correction Method

Trace elements tend to accumulate and bind to the clay/silt sediment fraction represented by particle sizes of less than 63 μm (Forstner and Wittmann 1983; Krumgalz et al. 1992).

Consequently, the heterogenous nature of the sediment in the St. Lawrence River makes it necessary to adjust trace element concentrations for the different particle size distributions at the various sampling stations in order to compare contaminant concentrations between stations if the effect of depositional environments are to be diminished and trace metal contaminant sources are to be inferred. There are several methods frequently used to correct for depositional environments and hence particle size differences between stations (Forstner and Wittmann 1983; Krumgalz et al. 1992).

The approach taken in this report was to normalize the anthropogenic trace metal results to a "conservative" element such as aluminum (i.e. an element that is not believed to be locally anthropogenic in origin). The ratio of the other metals to aluminum should remain constant across a gradient of particle sizes unless there is an enrichment of the other metal (Forstner 1990). Results from the 1994 survey showed some enrichment of aluminum in sediment samples within the study area (relative to the upstream reference area that was used for the comparison of sediment quality), likely due to anthropogenic atmospheric sources in Massena, (Richman 1996). To determine the appropriateness of this method of correction, aluminum concentrations measured in the surface sediment were compared with concentrations measured in the bottom of core samples (expected to represent historical aluminum concentrations), collected from the same locations. Using the 1994 data set, this type of comparison showed there was less than a 30% difference between surface and core bottom aluminum concentrations. This difference was within the range of variability for replicate core samples and suggests little change in aluminum concentrations over time. Accordingly, a correction for particle size was valid using this method although the potential exists to underestimate the local enrichment of other metals.

To further complement the results obtained from this correction method, the data were also corrected by normalizing metal concentrations at all stations to a fine particle (<63 μm) content of 74%. This involved the multiplication of the contaminant concentration at a site by 74% and then dividing by the actual percent silt plus clay for the station. This percentage (74%) was chosen since it was the median value for the particle size distribution for all samples collected in the study area. Another commonly used option would have been to normalize the concentration of contaminants to a fine particle content of 100% however this method has a tendency to produce inordinately high metal concentrations at stations high in sand.

To determine if the north channel downstream of the major point sources in Cornwall was contaminated with metals, a comparison was made between particle size corrected data from both this area and sediment data collected from Maitland, upstream of Cornwall. Results from a sediment survey in 1991 in the Maitland area where 48 stations were sampled along 11 transects were used for the comparison (Richman and Townsend 1997). The extensive (116 samples in total) 1991 survey concluded that the St. Lawrence River at Maitland was not severely contaminated with metals or nutrients and concentrations of contaminants at most sites were consistent with background stations making this an appropriate reference area. Details on sampling methods and the study area are provided in Richman and Townsend (1997).

Accordingly, all the metal data from Maitland were corrected for particle size by calculating the ratio of the concentration of a particular metal to the concentration of aluminum. The mean, and upper and lower 95% confidence intervals were calculated for Maitland and then compared with the normalized Cornwall data. All metal:Al ratios at Cornwall stations greater than the upper confidence limit for reference area metal:Al ratios were identified as being enriched with the particular parameter. The same procedure was followed using the correction method of normalizing the data to 74% fines.

Statistical Approach to Assess Temporal Changes in Sediment Contaminant Concentrations

Since the EC lab was responsible for the 1997 analysis and the MOE lab analysed the 1994 sediment samples, an inter-lab comparison of metals and particle size analysis was required in order to assess temporal trends. The inter-lab comparison was conducted through the splitting of 8 samples from different locations for analysis by each lab. A paired t-test was used to compare the metals data generated by the two laboratories. The results from the paired t-test were then used to identify the parameters appropriate for the “between year” comparison to determine if contaminant concentrations in sediment had changed over time.

The comparison of 1994 and 1997 data corrected for particle size from six stations was made with a paired t-test by grouping stations within a sampling year together to characterize the study area as a whole. This approach was necessary since there was only one sample at each station. Contaminant concentrations in 10 cm core samples were used for the comparison. Because of the small sample size overall (only six stations for the comparison), and the wide range in contaminant concentrations, a large change in concentration over time would be required to obtain a statistically significant difference between 1994 and 1997 data. As such the change in contaminant concentrations over time (using data that was both corrected and uncorrected for particle size), was reviewed by simple subtraction of concentrations between years to investigate general trends.

Another method to assess a change in contaminant concentrations over time was to compare concentrations of metals in the 1997 surface sediment samples (top 3 cm) with the concentrations in the 1997 core samples (top 10 cm) collected from the same stations using sampling sites from both the north and south side of the channel. This was based on the assumption that the top 3 cm of sediment consisted of more recent material than the top 10 cm. An analysis of covariance (ANCOVA) was used for the comparison between the sediment quality in these two types of samples. Particle size (% silt plus clay) was used as the covariate. The heterogeneity of slopes was tested for each parameter as a preliminary step using the SAS statistical package (SAS Institute Inc. 1988). The ANCOVA was performed for each parameter if the slopes generated for the surface samples and 10 cm core samples were not significantly different. The ANCOVA was performed on both log transformed and non-transformed data. Results from both tests were almost identical, as such results from the log transformed data were presented in this report.

Quality Assurance / Quality Control

Statistical analysis and results for QA/QC samples are provided in Appendix C.

RESULTS AND DISCUSSION

Patterns of Sediment Contamination

Data for sediment contaminant concentrations and particle size for stations sampled in the St. Lawrence River are summarized in Table 2 & 3.

All sediment samples (10 cm cores and 3 cm surface samples) collected along the Cornwall waterfront (north shore of the channel) exceeded the LEL for mercury and 46 % of the samples exceeded the SEL (Table 2, Figure 2). These results were consistent with previous sediment surveys which found that the north side of the channel was contaminated with mercury (MOE 1979; Kauss et al. 1988; Anderson 1991; Richman 1994, 1996). In contrast, all surface samples had mercury concentrations less than the LEL on the south side of the channel. These data also confirm results from previous surveys and indicate that this area was not contaminated with mercury (MOE 1979; Kauss et al. 1988; Anderson 1991). Median concentrations of mercury in core top samples (3.06 $\mu\text{g/g}$) were higher on the north shore compared with the south shore (0.13 $\mu\text{g/g}$) (median concentrations and range provided in Table 4).

In general throughout the study area concentrations of cadmium, nickel, chromium, copper, lead, zinc, total phosphorus and TOC were above the LEL. A comparison of metal sediment concentrations in samples (10 cm cores) collected from the north channel showed that median concentrations of cadmium (0.89 $\mu\text{g/g}$), nickel (24 $\mu\text{g/g}$) and TOC (2.5%) along the north shore were similar to concentrations detected on the south side of the channel (1.02 $\mu\text{g/g}$, 30.8 $\mu\text{g/g}$ and 2.5% respectively). Median concentrations of copper and lead were also similar on both sides of the channel but the range in concentrations for copper and lead was higher on the north shore (particularly downstream of Courtaulds) than the south shore (Table 4). This is consistent with data from previous surveys and the discharge patterns from local industries (MOE 1991; MOE 1992a&b). Median concentrations of zinc (186 $\mu\text{g/g}$) were higher on the north shore compared with the south shore (136 $\mu\text{g/g}$) as would be expected based on historical discharge patterns from Courtaulds (Kauss et al. 1988; Anderson 1991; Richman 1994, 1996) (Figure 3). Enrichment of zinc, lead and copper on the north shore downstream of Courtaulds was confirmed when data were corrected for particle size and compared with contaminant concentration upstream and on the south side of the channel.

All sediment samples from the north side of the channel (along the waterfront) were below the LEL for manganese and only one station had concentrations of iron above the LEL as compared with the south side of the channel where 57 % of the samples exceeded the LEL for manganese and all the samples exceeded the LEL for iron. Median concentrations of aluminum, iron, manganese, total phosphorus and chromium were higher on the south shore than the north. Once the sediment data were corrected for particle size there were no differences between both sides of the channel for any of these parameters.

TABLE 2: Metal, total phosphorus and TOC concentrations ($\mu\text{g/g}$ dry weight) in sediment samples (top 10 cm core sample) collected from the St. Lawrence River, 1997. (n=1). B - bottom 10 cm of the core. S - surface sample (top 3 cm).

| ENV.CAN STN # | AL $\mu\text{g/g}$ | CD $\mu\text{g/g}$ | CR $\mu\text{g/g}$ | CU $\mu\text{g/g}$ | FE $\mu\text{g/g}$ | HG $\mu\text{g/g}$ | MN $\mu\text{g/g}$ | NI $\mu\text{g/g}$ | PB $\mu\text{g/g}$ | ZN $\mu\text{g/g}$ | Total P $\mu\text{g/g}$ | TOC % |
|------------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|----------------------------|----------|
| CS105 | 19600 | 0.952 | 46.8 | 45.1 | 19300 | 1.67 | 295 | 26.8 | 63.3 | 600 | 761 | 2.926 |
| CS109 | 16500 | 1.170 | 44.0 | 55.4 | 16600 | 4.83 ** | 252 | 26.2 | 136.0 | 759 | 638 | 2.872 |
| CS109-S | 19000 | 0.929 | 49.1 | 55.6 | 18600 | 5.79 ** | 317 | 26.1 | 156.0 | 673 | 1020 | 2.753 |
| CS115 | 15400 | 1.450 | 41.8 | 57.9 | 16400 | 3.66 ** | 281 | 26.6 | 99.8 | 603 | 678 | 2.446 |
| CS115-S | 19400 | 1.080 | 43.7 | 57.6 | 19400 | 1.63 | 339 | 28.0 | 60.5 | 335 | 977 | 2.266 |
| CS117 | 11500 | 0.608 | 31.4 | 34.9 | 13300 | 2.00 ** | 245 | 19.9 | 83.5 | 279 | 549 | 2.353 |
| CS117-S | 18700 | 0.869 | 40.6 | 42.8 | 18600 | 1.35 | 333 | 25.8 | 60.8 | 270 | 965 | 2.049 |
| CS126 | 96300 | 0.720 | 23.8 | 26.6 | 12700 | 5.20 ** | 201 | 19.1 | 25.2 | 125 | 457 | 3.076 |
| CS126-S | 12200 | 0.424 | 28.5 | 21.8 | 18800 | 1.20 | 357 | 16.0 | 19.7 | 82 | 852 | 2.980 |
| CS127 | 13500 | 0.473 | 27.8 | 29.4 | 13600 | 4.32 ** | 280 | 13.8 | 53.6 | 152 | 548 | 1.681 |
| CS128 | 16800 | 0.637 | 32.7 | 33.6 | 17900 | 11.20 ** | 345 | 18.0 | 37.3 | 177 | 681 | 2.850 |
| CS128-S | 10300 | 0.359 | 25.1 | 21.1 | 15200 | 3.44 ** | 293 | 14.5 | 23.6 | 69 | 765 | 1.901 |
| CS131 | 17000 | 0.645 | 33.3 | 42.5 | 18300 | 19.50 ** | 347 | 17.8 | 40.6 | 186 | 659 | 4.037 |
| CS131-S | 14900 | 0.758 | 32.0 | 37.6 | 14700 | 14.70 ** | 288 | 24.9 | 33.7 | 251 | 853 | 2.737 |
| CS132 | 21800 | 0.949 | 43.8 | 60.9 | 17300 | 6.78 ** | 329 | 24.6 | 63.3 | 474 | 818 | 2.516 |
| CS135 | 19000 | 0.984 | 40.3 | 45.5 | 18500 | 3.75 ** | 331 | 22.9 | 58.8 | 234 | 660 | 2.108 |
| CS135-S | 12300 | 0.624 | 27.2 | 19.3 | 14300 | 1.22 | 288 | 21.1 | 27.2 | 117 | 758 | 1.753 |
| CS156 | 11400 | 0.490 | 26.2 | 16.9 | 14400 | 0.80 | 272 | 14.6 | 26.7 | 134 | 745 | 1.740 |
| CS164 | 16000 | 0.888 | 48.7 | 59.9 | 17700 | 3.09 ** | 292 | 24.9 | 136.0 | 669 | 847 | 2.454 |
| CS164-S | 14800 | 0.795 | 38.4 | 47.3 | 15900 | 1.98 | 291 | 25.2 | 100.0 | 395 | 841 | 2.822 |
| CS166 | 24400 | 1.220 | 47.5 | 43.2 | 23000 | 0.79 | 360 | 27.5 | 36.0 | 132 | 1150 | 3.236 |
| CS166-B | 27000 | 1.290 | 53.1 | 46.1 | 23500 | 0.12 | 328 | 32.8 | 44.2 | 142 | 1140 | 3.263 |
| CS166-S | 23400 | 1.120 | 45.3 | 43.4 | 22200 | 1.06 | 335 | 30.2 | 33.7 | 133 | 986 | 3.491 |
| CS167 | 16800 | 0.910 | 37.0 | 39.7 | 16300 | 1.19 | 265 | 29.9 | 36.6 | 107 | 867 | 2.970 |
| CS167-B | 24700 | 1.420 | 44.0 | 64.4 | 15000 | 4.63 ** | 247 | 38.8 | 58.3 | 158 | 1120 | 6.745 |
| CS167-S | 12200 | 0.687 | 26.7 | 23.6 | 13700 | 1.72 | 239 | 19.5 | 21.9 | 76.9 | 653 | 2.370 |
| CS168 | 17000 | 0.877 | 37.5 | 31.7 | 17000 | 1.71 | 290 | 24.0 | 38.0 | 102 | 947 | 1.813 |
| CS168-B | 25600 | 1.590 | 51.5 | 75.1 | 17000 | 13.70 ** | 286 | 35.9 | 73.6 | 187 | 1090 | 5.818 |
| CS168-S | 11800 | 0.754 | 26.8 | 21.2 | 13100 | 0.70 | 222 | 20.8 | 24.7 | 80.7 | 724 | 2.178 |
| Lowest Effect Level | 0.6 | 26 | 16 | 2% | 0.2 | 460 | 16 | 31 | 120 | 600 | 1 | |
| Severe Effect Level ** | 10 | 110 | 110 | 4% | 2 | 1100 | 75 | 250 | 820 | 2000 | 10 | |

TABLE 2 continued:

| ENV.CAN STM # | AL ug/g | CD ug/g | CR ug/g | CU ug/g | FE ug/g | HG ug/g | MN ug/g | NI ug/g | PB ug/g | ZN ug/g | Total P ug/g | TOC % |
|------------------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|------------|-----------------|----------|
| CS171 | 18600 | 0.690 | 36.5 | 31.3 | 19100 | 0.44 | 359 | 20.6 | 29.6 | 118 | 918 | 2.016 |
| CS172 | 31200 | 1.120 | 58.8 | 46.1 | 26600 | 0.62 | 439 | 34.1 | 44.0 | 350 | 1120 | 3.340 |
| CS173 | 32700 | 1.210 | 58.4 | 46.1 | 28900 | 0.13 | 489 | 34.8 | 39.3 | 155 | 1150 | 3.130 |
| CS175 | 29200 | 1.020 | 52.4 | 39.7 | 26500 | 0.08 | 464 | 30.4 | 37.5 | 136 | 1080 | 2.532 |
| CS175-B | 23000 | 1.380 | 48.5 | 36.4 | 21600 | 0.24 | 359 | 30.8 | 46.8 | 131 | 931 | 2.246 |
| CS176 | 33300 | 1.150 | 59.1 | 46.3 | 30000 | 0.15 | 592 | 35.4 | 39.0 | 156 | 1180 | 3.099 |
| CS176-S | 27500 | 1.100 | 48.7 | 41.5 | 26200 | 0.15 | 665 | 38.4 | 29.5 | 136 | 1110 | 3.229 |
| CS177 | 28600 | 0.976 | 49.5 | 38.1 | 26000 | 0.12 | 495 | 30.8 | 32.8 | 132 | 1020 | 2.599 |
| CS179 | 28200 | 1.020 | 50.6 | 38.9 | 26500 | 0.14 | 488 | 31.5 | 35.2 | 136 | 1090 | 2.227 |
| CS-179-B | 20400 | 1.070 | 41.5 | 30.0 | 20300 | 0.16 | 344 | 26.9 | 36.2 | 108 | 909 | 1.945 |
| CS-179-S | 23900 | 1.040 | 42.9 | 34.0 | 22900 | 0.12 | 445 | 33.7 | 26.5 | 120 | 956 | 3.079 |
| CS181 | 25500 | 1.010 | 47.9 | 35.6 | 24300 | 0.13 | 435 | 29.9 | 33.2 | 126 | 996 | 2.030 |
| CS-181-B | 22900 | 1.290 | 47.7 | 35.9 | 22300 | 0.20 | 366 | 31.2 | 43.3 | 124 | 952 | 2.141 |
| CS-181-S | 21700 | 0.916 | 39.7 | 30.6 | 20800 | 0.13 | 386 | 28.4 | 27.0 | 110 | 698 | 2.528 |
| CS182 | 24300 | 0.901 | 43.8 | 31.8 | 23500 | 0.12 | 492 | 28.4 | 27.2 | 117 | 1090 | 2.507 |
| CS-182-S | 23500 | 0.939 | 42.4 | 33.4 | 23000 | 0.12 | 513 | 32.3 | 26.3 | 116 | 986 | 2.827 |
| Lowest Effect Level | 0.6 | 26 | 16 | 2% | 0.2 | 460 | 16 | 31 | 120 | 600 | 1 | |
| Severe Effect Level ** | 10 | 110 | 110 | 4% | 2 | 1100 | 75 | 250 | 820 | 2000 | 10 | |

TABLE 3: Particle size data for sediment samples (top 10 cm core sample) collected from the St. Lawrence River, 1997. (n=1)
 B - bottom 10 cm of the core. S - surface sample (top 3 cm).

| Station Number | Percent Sand | Percent Silt | Percent Clay | Station Number | Percent Sand | Percent Silt | Percent Clay |
|----------------|--------------|--------------|--------------|----------------|--------------|--------------|--------------|
| CS105 | 15.7 | 52.4 | 31.9 | CS167-S | 36.7 | 45.4 | 17.9 |
| CS109 | 14.2 | 55.9 | 29.9 | CS167 | 27.9 | 51.6 | 20.5 |
| CS109-S | 9.0 | 68.6 | 22.4 | CS167-B | 8.7 | 60.6 | 30.6 |
| CS115 | 17.6 | 60.9 | 21.5 | CS168-S | 45.0 | 38.7 | 16.2 |
| CS115-S | 33.0 | 42.8 | 24.1 | CS168 | 42.1 | 42.1 | 15.9 |
| CS117-S | 42.2 | 35.2 | 22.5 | CS168-B | 3.7 | 62.4 | 34.0 |
| CS117 | 51.2 | 33.3 | 15.5 | CS171 | 35.2 | 43.6 | 21.1 |
| CS126-S | 55.6 | 28.4 | 16.0 | CS172 | 9.9 | 59.4 | 30.6 |
| CS126 | 58.3 | 30.5* | 10.5* | CS173 | 5.4 | 66.0 | 28.6 |
| CS127 | 68.6 | 18.2 | 13.2 | CS175 | 12.9 | 59.1 | 28.0 |
| CS128-S | 65.8 | 18.8 | 15.4 | CS175-B | 13.7 | 64.8 | 21.5 |
| CS128 | 39.8 | 42.3 | 16.8 | CS176-S | 3.6 | 61.8 | 34.6 |
| CS131-S | 25.9 | 52.6 | 21.5 | CS176 | 6.2 | 65.0 | 28.8 |
| CS131 | 35.7 | 43.1 | 21.2 | CS177 | 13.5 | 58.4 | 28.2 |
| CS132 | 39.0 | 38.0 | 23.0 | CS179-S | 11.4 | 59.1 | 29.5 |
| CS135-S | 38.6 | 39.5 | 21.8 | CS179 | 14.2 | 57.7 | 28.1 |
| CS135 | 33.2 | 44.9 | 21.9 | CS179-B | 22.1 | 55.0 | 22.9 |
| CS156 | 48.6 | 31.7 | 19.7 | CS181-S | 11.8 | 60.1 | 28.1 |
| CS164-S | 16.5 | 60.0 | 23.5 | CS181 | 11.5 | 62.2 | 26.3 |
| CS164 | 29.0 | 53.6 | 17.4 | CS181-B | 16.2 | 59.6 | 24.2 |
| CS166-S | 9.3 | 66.4 | 24.3 | CS182 | 14.0 | 56.1 | 29.9 |
| CS166 | 7.6 | 66.2 | 26.2 | CS182-S | 9.8 | 57.1 | 33.1 |
| CS166-B | 6.1 | 65.4 | 28.5 | | | | |

* estimated value-lab reported 41% for silt+clay

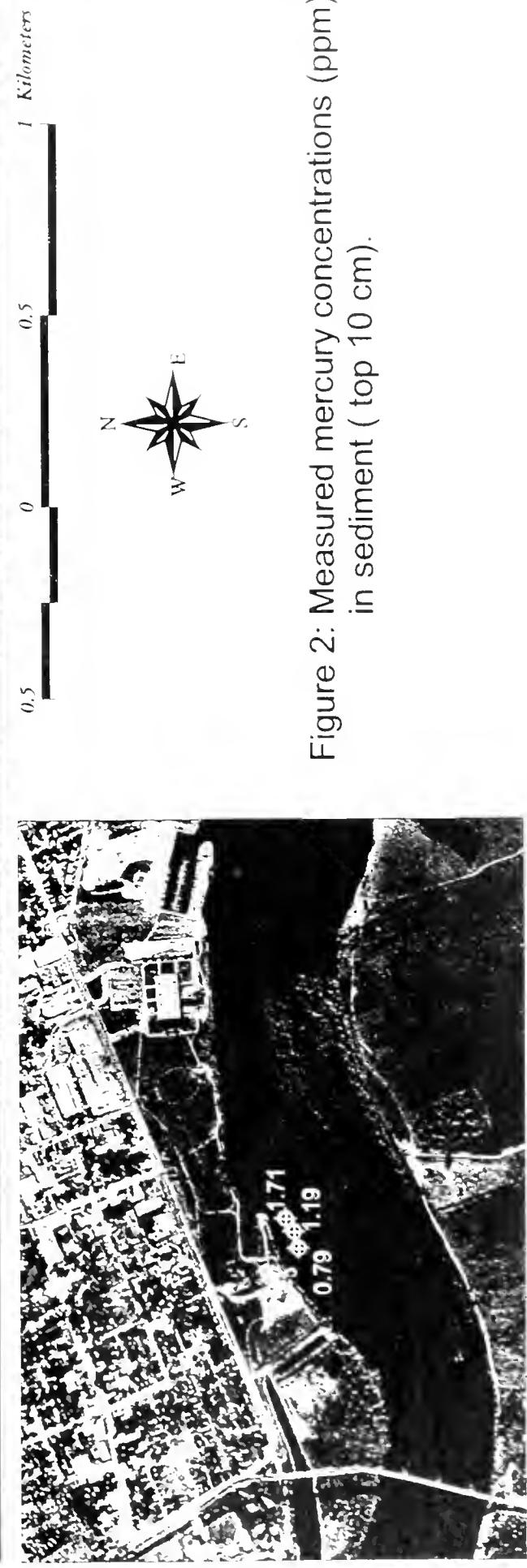
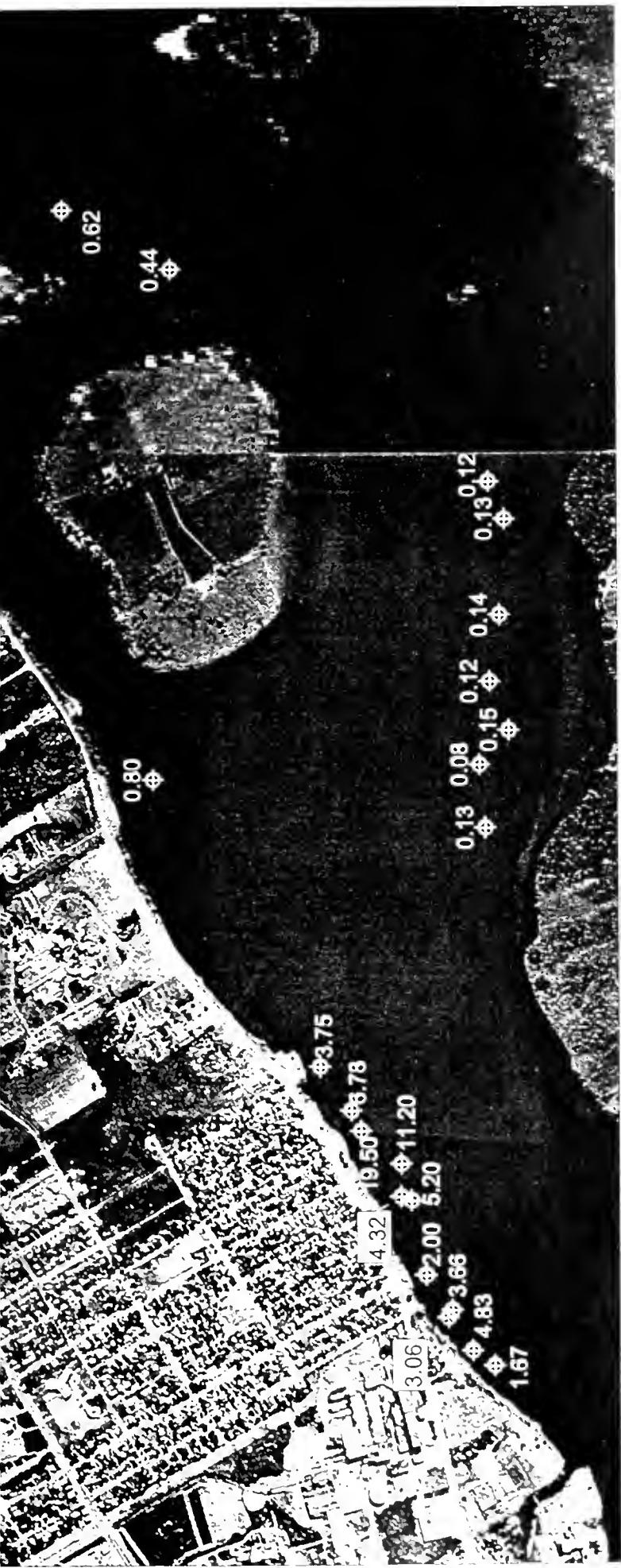


Figure 2: Measured mercury concentrations (ppm) in sediment (top 10 cm).

TABLE 4: Median metal, total phosphorus and TOC concentrations ($\mu\text{g/g}$ dry weight) in sediment samples (top 10 cm core sample and 3 cm surface samples) collected from the St. Lawrence River, 1997.

| Parameters | Median, Minimum and Maximum Concentrations of Contaminants in Sediment | | | | | | | | | | | |
|------------------|--|-------|-------|-----------------|-------------|-------|----------------|-------|-------|-----------------|-------|-------|
| | North Shore | | | | South Shore | | | | | | | |
| | Surface (3 cm) | min | max | Core (10 cm) | min | max | Core (3 cm) | min | max | Core (10 cm) | min | max |
| Aluminum | 14800 | 10300 | 23400 | 16800 | 9630 | 31200 | 23700 | 21700 | 27500 | 28600 | 24300 | 33300 |
| Cadmium | 0.758 | 0.359 | 1.120 | 0.888 | 0.473 | 1.450 | 0.990 | 0.916 | 1.100 | 1.020 | 0.901 | 1.210 |
| Chromium | 32.0 | 25.1 | 49.1 | 37.5 | 23.8 | 58.8 | 42.7 | 39.7 | 48.7 | 50.6 | 43.8 | 59.1 |
| Copper | 37.6 | 19.3 | 57.6 | 42.5 | 16.9 | 60.9 | 33.7 | 30.6 | 41.5 | 38.9 | 31.8 | 46.3 |
| Iron | 15900 | 13100 | 22200 | 17300 | 12700 | 26600 | 22950 | 20800 | 26200 | 26500 | 23500 | 30000 |
| Mercury | 1.63 | 0.70 | 14.70 | 3.06 | 0.44 | 19.50 | 0.13 | 0.12 | 0.15 | 0.13 | 0.08 | 0.15 |
| Manganese | 293 | 222 | 357 | 292 | 201 | 439 | 479 | 386 | 665 | 489 | 435 | 592 |
| Nickel | 24.9 | 14.5 | 30.2 | 24.0 | 13.8 | 34.1 | 33.0 | 28.4 | 38.4 | 30.8 | 28.4 | 35.4 |
| Total Phosphorus | 852 | 653 | 1020 | 745 | 457 | 1150 | 971 | 698 | 1110 | 1090 | 996 | 1180 |
| Lead | 33.7 | 19.7 | 156.0 | 44.0 | 25.2 | 136.0 | 26.8 | 26.3 | 29.5 | 35.2 | 27.2 | 39.3 |
| Zinc | 133 | 69 | 673 | 186 | 102 | 759 | 118 | 110 | 136 | 136 | 117 | 156 |
| TOC (%) | 2.370 | 1.753 | 3.491 | 2.516 | 1.681 | 4.037 | 2.953 | 2.528 | 3.229 | 2.532 | 2.030 | 3.130 |

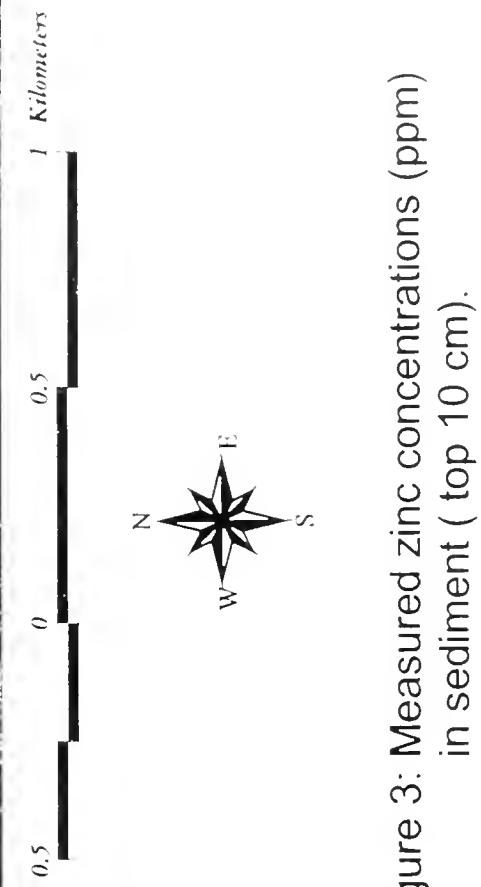
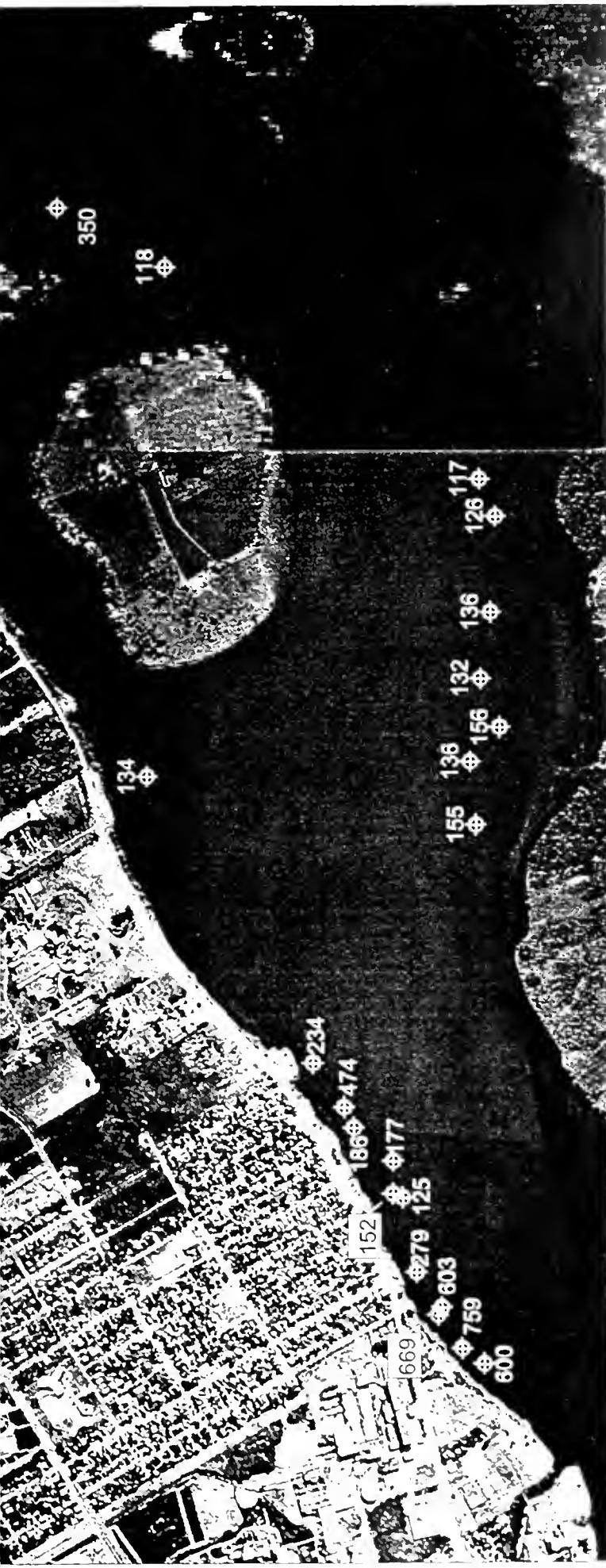


Figure 3: Measured zinc concentrations (ppm) in sediment (top 10 cm).

The surface samples (3 cm) collected from both sides of the channel showed a similar pattern of sediment quality as discussed above for the 10 cm core samples.

The principal components analysis (PCA) was performed on log transformed data using surface samples (3 cm) and core samples (top 10 cm). The first axis grouped all variables together with the exception of lead, copper and zinc which all subsequently ranked high on the second axis (Table 5). Mercury was grouped on both the first and second axis however it was negatively correlated with the other parameters on the first axis and positively correlated with lead, copper and zinc on the second axis. The second axis was dominated by the variables associated with the former local discharge of contaminants from Courtaulds. These parameters were correlated with one another and stations with high concentrations of Hg, Zn, Cu, and Pb grouped together and were all located downstream, or in the vicinity of the Courtaulds shore based outfalls (Figure 4).

The stations on the south shore of the channel where concentrations of the parameters were low were also all grouped together in the lower right side of the graph. Figure 4 shows station 167 and 168 (downstream of the Domtar/ICI diffuser) grouped with stations that were high in Hg but with relatively low concentrations of Zn, Pb, and Cu. The first axis explained 59% of the variation in the data. By including the second axis 81% of the variability was explained.

The PCA using only percent silt produced results which were the same as the original analysis which used percent silt, clay and sand ensuring that the outcome of the first PCA was not artificially driven by overweighting the analysis with particle size variables (Appendix D). The PCA was again repeated using only data from the north shore and the results and overall conclusions were the same as described above. As well, the PCA using only core data also resulted in the same conclusions.

Particle Size Corrections of Sediment Contaminant Data

A comparison of particle size corrected data from Cornwall with data from Maitland shows that the sediment downstream of the Cornwall point sources was not enriched with cadmium, chromium, iron, manganese and nickel. For these parameters, the ratios (metal:Al) generated for surface and core samples collected from both the north and south side of the channel were all less than the lower 95% confidence interval or were within the confidence bands generated for the Maitland data. When the data were corrected by normalizing metal concentrations at all stations to a fine particle (<63 μm) content of 74 % the results showed the same pattern. All data for the comparisons between the study area and reference area are provided in Appendix B(1), B(2) & B(3).

The copper, lead, mercury and zinc ratios for samples collected from the south side of the north channel were all less than the lower 95% confidence interval generated for the upstream reference area confirming previous observations that there was no enrichment of these parameters along the south shore. The corrected concentrations for these parameters calculated

TABLE 5: Component loadings and percent of total variance explained for the PCA of sediment quality in the St. Lawrence River, 1997.

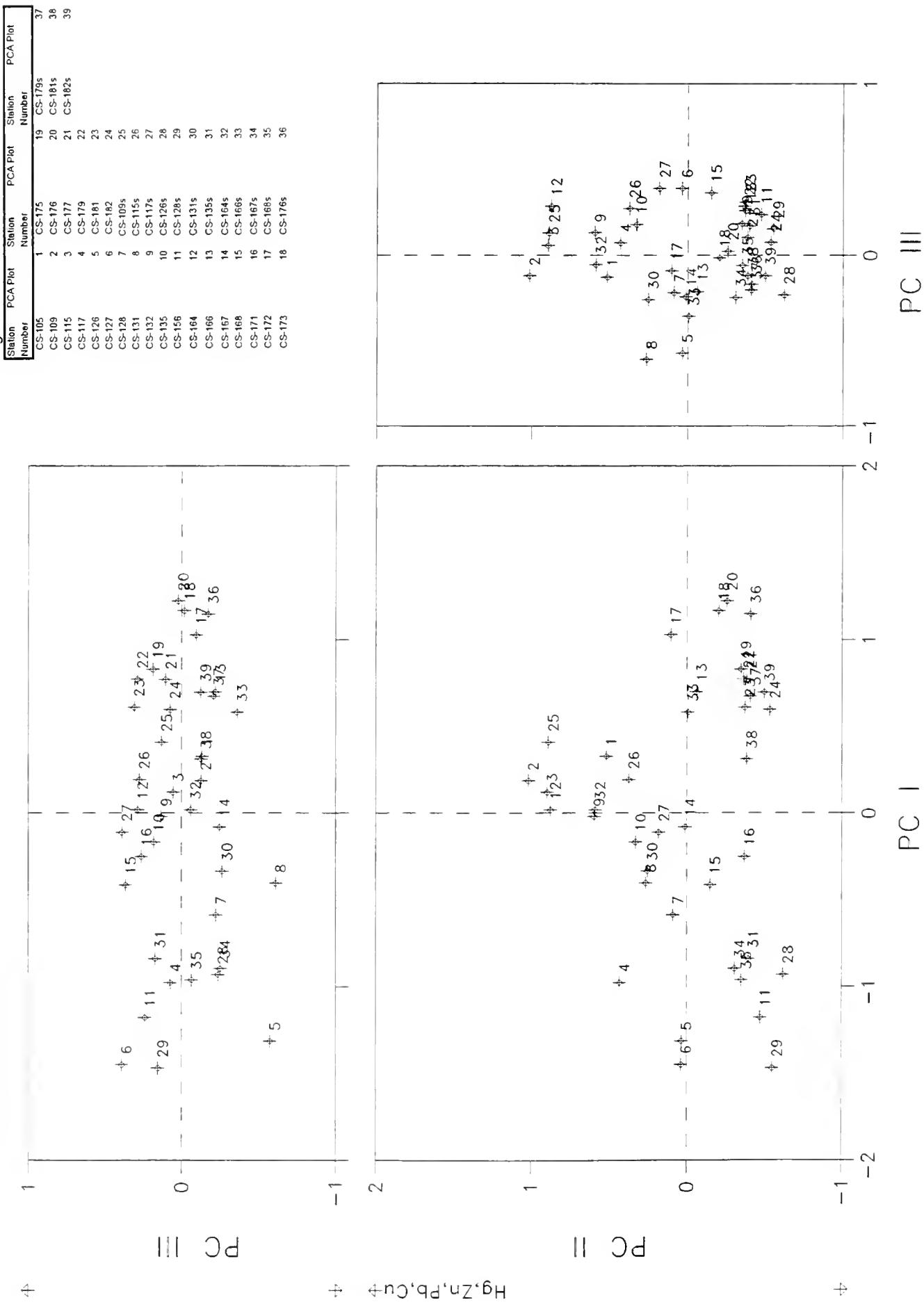
Log Transformed Data

| | PC I | PC II | PC III |
|------------|--------|--------|--------|
| Aluminum | 0.944 | -0.141 | 0.1 |
| Cadmium | 0.842 | 0.3 | -0.026 |
| Chromium | 0.948 | 0.19 | 0.175 |
| Copper | 0.621 | 0.721 | 0.044 |
| Iron | 0.904 | -0.291 | 0.06 |
| Mercury | -0.612 | 0.653 | -0.207 |
| Manganese | 0.79 | -0.431 | 0.104 |
| Nickel | 0.936 | 0.018 | -0.028 |
| Phosphorus | 0.785 | -0.295 | 0.23 |
| Lead | 0.135 | 0.922 | 0.271 |
| Zinc | 0.208 | 0.923 | 0.13 |
| TOC | 0.496 | 0.176 | -0.8 |
| Sand | -0.939 | -0.058 | 0.136 |
| Silt | 0.891 | 0.12 | -0.113 |
| Clay | 0.859 | -0.11 | -0.14 |

Percent of total variance explained

59% 81% 87%

Figure 4



<- Sand, Hg

Al, Cd, Cr, Fe, Mn, Ni, Clay

PC I

PC II

by normalizing metal concentrations at all stations to a fine particle (<63 μm) content of 74 % provided the same conclusions supporting the results from the aluminum normalization method.

Core samples collected along the Cornwall waterfront at stations CS109, CS115, CS117 and CS164 had Cu:Al and Pb:Al ratios that were greater than the upper 95% confidence interval generated for Maitland indicating enrichment (Figure 5). For copper, this also occurred at station CS126 and CS132 and for lead this also occurred at station CS127. The results suggest enrichment of copper and lead at these stations. For surface samples (top 3 cm) stations CS109, CS115 and CS164 showed copper enrichment relative to Maitland while only station CS109 and CS164 showed lead enrichment. These stations were all located downstream of Courtaulds which has been associated with the discharge of these parameters.

All the core samples collected on the north shore had higher ratios of Zn:Al than the 95% upper confidence limit generated for the upstream reference area with the exception of stations CS166, CS167 and CS168 located downstream of the Domtar/ICI diffuser but upstream of Courtaulds. This is consistent with patterns of zinc discharge which was primarily from Courtaulds. Ratios of Zn:Al were approximately two to six times higher downstream of Courtaulds when compared to ratios generated for stations located upstream of the facility. The ratios calculated for the surface samples (3 cm) generally showed the same pattern.

Corrected concentrations of zinc and lead calculated by normalizing metal concentrations at all stations to a fine particle (<63 μm) content of 74% showed the same pattern of contamination as the aluminum normalizing approach, with only one exception (station CS132). This method did not suggest copper enrichment relative to the upstream reference area, although qualitatively stations along the north shore were enriched relative to the south shore using corrected and uncorrected data. This is effectively an artifact of the correction method since some sampling stations in Maitland were greater than 90% sand. When corrected to a fine particle size content of 74% the extrapolated copper concentrations become exceedingly large making it more difficult to show enrichment.

All samples (surface and 10 cm core samples) along the north shore had higher Hg:Al ratios than the 95% upper confidence limits generated for Maitland, suggesting enrichment of all stations downstream of Cornwall point sources. This is consistent with local mercury discharge patterns. Ratios for the core samples ranged from 2 to 133 times higher than the upper confidence limit for the Maitland area with most ratios greater than 20 times higher. For surface samples, the ratios of Hg:Al were 5 to 114 times higher in Cornwall than the 95% upper confidence limit generated for Maitland. The same results applied to the particle size correction method.

Contaminants in Core Top Samples Compared With Core Bottom Samples

Metal, total phosphorus and TOC concentrations in the bottom of the core samples collected from the three stations downstream of Domtar/ICI and the three stations from the south shore deposit were compared with the 10 cm core top samples and 3 cm surface samples from the same



Figure 5: Local sediment enrichment of copper and lead.

locations to determine if there had been changes in sediment quality over time and to confirm whether these locations were historically contaminated with mercury. Concentrations of all parameters at station CS-166 (Downstream of Domtar/ICI) and at station CS175, CS179 and CS181 (south side deposit) in the bottom of the cores were similar to concentrations found in the top 10 cm core samples and surface samples for all metals, total phosphorus and TOC (Table 2). This suggests little change in sediment quality over time at these four stations based on the assumption that the samples collected from the bottom of the cores reflect historical conditions relative to the sediment located at the top of the cores. The particle size distribution (percent silt, clay and sand), in the top of the cores for these four stations was almost identical to the particle size distribution in samples collected from the bottom of the cores which strengthens the comparison of metal concentrations between the two sample types.

At station CS167 and CS168 (downstream of Domtar/ICI), sediment concentrations for many of the parameters were higher in the core bottom samples when compared with the 10 cm core top samples and 3 cm surface samples, however, this was likely due to differences in particle size distributions between the samples. The bottom of the cores were high in percentage clay plus silt (91.2% and 96.4% respectively) while the core top samples were much higher in percentage sand (Table 3). The percentage clay plus silt in the top of the cores was only 72.1% for station CS167 and 58% for CS168. The 3 cm surface samples at these two stations had similar metal concentrations as the 10 cm core top samples and similar particle size distributions (Table 2 & 3).

The metal concentrations in the samples from these two stations were corrected by normalizing the trace metals to aluminum concentrations. The ratios for all the parameters (with the exception of mercury), were similar between the core top samples and the core bottom samples verifying that the discrepancies in metal concentrations were likely due to particle size differences (Appendix B1).

However for mercury the correction method generated ratios that were three and five times higher in the bottom samples than the top samples respectively for station CS167 and CS168. This suggest that the high concentrations of mercury in the bottom samples at stations CS-167 and CS-168 were due to enrichment and cannot be due to particle size differences alone. These high concentrations most likely were due to upstream discharges of mercury from the Fly Creek sewer which historically discharged effluent from Domtar and ICI. Although station CS166 was close to stations CS167 and CS168 the data suggest it was not impacted by the upstream source since mercury concentrations were low in core top and bottom samples. More samples collected from this area would provide a better indication of the areal extent of impact by the upstream point source. Previous surveys show a patchy distribution in the area. However, these limited data support results from previous studies which showed that this area was contaminated. Historical data show that mercury concentrations were as high as 18.2 $\mu\text{g/g}$ in 1975 and 19.8 $\mu\text{g/g}$ in 1979 in the vicinity of the 1997 stations (MOE 1979; Kauss et al. 1988).

Comparison of 1997 Data with 1994 Data

Split samples for interlaboratory comparisons between MOE and EC for metal and particle size analysis were compared using a paired t-test. The paired t-test showed there was no significant difference in trace metal data between the two laboratories for iron, chromium, mercury, manganese, nickel and lead ($p>0.05$). As well, there was no significant difference between the two laboratories for particle size analysis. The EC laboratory consistently reported higher aluminum values than the MOE lab ($p<0.001$), while the MOE lab reported higher cadmium ($p<0.001$), copper ($p<0.01$) and zinc ($p<0.02$) concentrations. Results of the statistical analysis are presented in Table 6 (raw data are provided in Appendix E). A comparison of sediment data between years to determine changes in sediment quality over time was therefore limited to parameters with concentrations that were not found to be significantly different between laboratories.

The effect of these results on the particle size correction methods and subsequent comparison of the sediment quality in the Cornwall area with the upstream reference area which had sediment that was analyzed by the MOE laboratory, actually underestimates the contamination in Cornwall. The identification of sites that were enriched with Hg, Zn, Cu and Pb was conservative using either method of particle size correction with both correction methods likely underestimating the number of sites that were enriched.

Since previous surveys suggest that the sediment quality in the study area was heterogenous (MOE 1979; Richman 1996), direct comparison of data on a station by station basis between years required that the stations be located close to one another for the comparison to be relevant. The 1994 QA/QC data using triplicate core samples from six stations showed that within station variability increased when the distance between replicate cores increased from 1 to 7 meters. The 1997 QA/QC data (Appendix C) for within station variability showed good agreement in concentrations for samples collected about 7 meters apart but variability increased as distance between samples increased. Based on the northing and easting recorded in the field in 1997, stations CS105, CS109, CS117, CS127, CS135 and CS164 were not considered to be the same as their counterparts in the 1994 survey. The remaining stations were close (<8 meters) to the original stations in 1994 allowing for better data comparisons (northing and easting data provided in Appendix A).

Accordingly, results from the 1997 survey for stations CS115, CS126, CS128, CS131, CS132 and CS156 were compared with results from the same locations collected in 1994 using a paired t-test to determine if there was a change in sediment quality. Using 1994 and 1997 data corrected for particle size differences (by normalizing the contaminant concentrations to a particle content of 74% fines) there was no significant difference in sediment concentrations between 1994 and 1997 for mercury, chromium, nickel, iron and manganese. There was a statistical difference in concentrations between years for lead ($p<0.01$), phosphorus ($p<0.001$) and TOC ($p<0.01$) (Table 7).

**TABLE 6: Paired t-Test for a comparison between analytical results generated by the Ministry of Environment (MOE) and Environment Canada (EC).
Sediment samples collected from the St. Lawrence River, 1997 were split and analysed by the two laboratories.**

| Aluminum | | EC | | MOE | | Iron | | EC | | MOE | | Chromium | | EC | |
|------------------------------|------------|--------------|--|------------------------------|------------|-----------|--|------------------------------|---------|---------|--|----------|------------------------------|---------|---------|
| Mean | 11150 | 17462 | | Mean | 17250 | 17500 | | Mean | 35.8 | 38.6 | | | Mean | 35.8 | 38.6 |
| Variance | 4788571.43 | 1136553535.7 | | Variance | 6785714.29 | 4985714.3 | | Variance | 45.93 | 38.10 | | | Variance | 45.93 | 38.10 |
| Observations | 8 | 8 | | Observations | 8 | 8 | | Observations | 8 | 8 | | | Observations | 8 | 8 |
| Pearson Correlation | 0.868 | | | Pearson Correlation | 0.865 | | | Pearson Correlation | | | | | Pearson Correlation | 0.776 | |
| Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | | Hypothesized Mean Difference | 0 | |
| df | | | | df | | | | df | | | | | df | | |
| t Stat | | | | t Stat | | | | t Stat | | | | | t Stat | | |
| P(T \leq t) one-tail | | | | P(T \leq t) one-tail | | | | P(T \leq t) one-tail | | | | | P(T \leq t) one-tail | | |
| t Critical one-tail | | | | t Critical one-tail | | | | t Critical one-tail | | | | | t Critical one-tail | | |
| P(T \leq t) two-tail | | | | P(T \leq t) two-tail | | | | P(T \leq t) two-tail | | | | | P(T \leq t) two-tail | | |
| t Critical two-tail | | | | t Critical two-tail | | | | t Critical two-tail | | | | | t Critical two-tail | | |
| Cadmium | | EC | | MOE | | EC | | MOE | | EC | | Nickel | | EC | |
| Mean | 1.263 | 0.880 | | Mean | 2.22 | 2.86 | | Mean | 24.5 | 25.5 | | | Mean | 24.5 | 25.5 |
| Variance | 0.0427 | 0.0431 | | Variance | 5.009 | 10.926 | | Variance | 6.57 | 15.56 | | | Variance | 6.57 | 15.56 |
| Observations | 8 | 8 | | Observations | 8 | 8 | | Observations | 8 | 8 | | | Observations | 8 | 8 |
| Pearson Correlation | 0.842 | | | Pearson Correlation | 0.971 | | | Pearson Correlation | | | | | Pearson Correlation | 0.696 | |
| Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | | Hypothesized Mean Difference | 0 | |
| df | | | | df | | | | df | | | | | df | | |
| t Stat | | | | t Stat | | | | t Stat | | | | | t Stat | | |
| P(T \leq t) one-tail | | | | P(T \leq t) one-tail | | | | P(T \leq t) one-tail | | | | | P(T \leq t) one-tail | | |
| t Critical one-tail | | | | t Critical one-tail | | | | t Critical one-tail | | | | | t Critical one-tail | | |
| P(T \leq t) two-tail | | | | P(T \leq t) two-tail | | | | P(T \leq t) two-tail | | | | | P(T \leq t) two-tail | | |
| t Critical two-tail | | | | t Critical two-tail | | | | t Critical two-tail | | | | | t Critical two-tail | | |
| Copper | | EC | | MOE | | EC | | MOE | | EC | | Lead | | EC | |
| Mean | 47.3 | 39.8 | | Mean | 287.5 | 297.125 | | Mean | 58.5 | 50.8 | | | Mean | 58.5 | 50.8 |
| Variance | 235.64 | 72.09 | | Variance | 2678.57 | 1547.55 | | Variance | 2237.14 | 1355.35 | | | Variance | 2237.14 | 1355.35 |
| Observations | 8 | 8 | | Observations | 8 | 8 | | Observations | 8 | 8 | | | Observations | 8 | 8 |
| Pearson Correlation | 0.988 | | | Pearson Correlation | 0.863 | | | Pearson Correlation | | | | | Pearson Correlation | 0.995 | |
| Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | | Hypothesized Mean Difference | 0 | |
| df | | | | df | | | | df | | | | | df | | |
| t Stat | | | | t Stat | | | | t Stat | | | | | t Stat | | |
| P(T \leq t) one-tail | | | | P(T \leq t) one-tail | | | | P(T \leq t) one-tail | | | | | P(T \leq t) one-tail | | |
| t Critical one-tail | | | | t Critical one-tail | | | | t Critical one-tail | | | | | t Critical one-tail | | |
| P(T \leq t) two-tail | | | | P(T \leq t) two-tail | | | | P(T \leq t) two-tail | | | | | P(T \leq t) two-tail | | |
| t Critical two-tail | | | | t Critical two-tail | | | | t Critical two-tail | | | | | t Critical two-tail | | |

Table 6 continued

| Zinc | | MOE | | EC | | Sand-Sum3 | | MOE | | EC | |
|------------------------------|--|----------|-----|----------|-----|------------------------------|--|--------|--|---------|--|
| Mean | | | 323 | | 267 | Mean | | 35.6 | | 33.6 | |
| Variance | | 77535.71 | | 48803.27 | | Variance | | 204.84 | | 298.37 | |
| Observations | | 8 | | 8 | | Observations | | 8 | | 8 | |
| Pearson Correlation | | 0.997 | | | | Pearson Correlation | | 0.242 | | | |
| Hypothesized Mean Difference | | 0 | | | | Hypothesized Mean Difference | | 0 | | | |
| df | | 7 | | | | df | | 7 | | | |
| t Stat | | 2.601 | | | | t Stat | | 0.298 | | | |
| P(T<=t) one-tail | | 0.018 | | | | P(T<=t) one-tail | | 0.387 | | | |
| t Critical one-tail | | 1.895 | | | | t Critical one-tail | | 1.895 | | | |
| P(T<=t) two-tail | | 0.035 | | | | P(T<=t) two-tail | | 0.774 | | | |
| t Critical two-tail | | 2.365 | | | | t Critical two-tail | | 2.365 | | | |
| Clay-Sum1 | | MOE | | EC | | Silt-Sum2 | | MOE | | EC | |
| Mean | | 17.0 | | 18.7 | | Mean | | 47.1 | | 47.6 | |
| Variance | | 25.271 | | 20.067 | | Variance | | 94.696 | | 188.623 | |
| Observations | | 8 | | 8 | | Observations | | 8 | | 8 | |
| Pearson Correlation | | 0.489 | | | | Pearson Correlation | | 0.189 | | | |
| Hypothesized Mean Difference | | 0 | | | | Hypothesized Mean Difference | | 0 | | | |
| df | | 7 | | | | df | | 7 | | | |
| t Stat | | -1.903 | | | | t Stat | | -0.081 | | | |
| P(T<=t) one-tail | | 0.175 | | | | P(T<=t) one-tail | | 0.469 | | | |
| t Critical one-tail | | 1.895 | | | | t Critical one-tail | | 1.895 | | | |
| P(T<=t) two-tail | | 0.349 | | | | P(T<=t) two-tail | | 0.938 | | | |
| t Critical two-tail | | 2.365 | | | | t Critical two-tail | | 2.365 | | | |

TABLE 7: Paired t-Test for a comparison between mean surface sediment concentrations (10 cm core samples) at selected stations in 1994 with concentrations measured in 1997. All data were corrected for particle size.

| Chromium | | 1997 | 1994 | Manganese | | 1997 | 1994 | Total Phosphorus | | 1997 | 1994 |
|------------------------------|-----------|------------|------|------------------------------|--------|--------|------|------------------------------|---------|---------|------|
| Mean | 41.8 | 43.6 | | Mean | 373 | 334 | | Mean | 853 | 1475 | |
| Variance | 35.58 | 40.13 | | Variance | 3982.8 | 3056.4 | | Variance | 27462.9 | 44483.3 | |
| Observations | 6 | 6 | | Observations | 6 | 6 | | Observations | 6 | 6 | |
| Pearson Correlation | 0.056 | | | Pearson Correlation | 0.407 | | | Pearson Correlation | 0.600 | | |
| Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | |
| df | 5 | | | df | 5 | | | df | 5 | | |
| t Stat | -0.523 | | | t Stat | 1.475 | | | t Stat | -8.800 | | |
| P(T<=t) one-tail | 0.312 | | | P(T<=t) one-tail | 0.100 | | | P(T<=t) one-tail | 0.0002 | | |
| t Critical one-tail | 2.015 | | | t Critical one-tail | 2.015 | | | t Critical one-tail | 2.015 | | |
| P(T<=t) two-tail | 0.624 | | | P(T<=t) two-tail | 0.200 | | | P(T<=t) two-tail | 0.0003 | | |
| t Critical two-tail | 2.571 | | | t Critical two-tail | 2.571 | | | t Critical two-tail | 2.571 | | |
| Iron | | 1997 | 1994 | Nickel | | 1997 | 1994 | TOC | | 1997 | 1994 |
| Mean | 20501 | 20103 | | Mean | 25.4 | 30.4 | | Mean | 3.590 | 5.184 | |
| Variance | 8711476.9 | 90008297.1 | | Variance | 31.26 | 23.90 | | Variance | 1.6704 | 0.4412 | |
| Observations | 6 | 6 | | Observations | 6 | 6 | | Observations | 6 | 6 | |
| Pearson Correlation | -0.048 | | | Pearson Correlation | -0.566 | | | Pearson Correlation | 0.626 | | |
| Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | |
| df | 5 | | | df | 5 | | | df | 5 | | |
| t Stat | 0.227 | | | t Stat | -1.306 | | | t Stat | -3.835 | | |
| P(T<=t) one-tail | 0.415 | | | P(T<=t) one-tail | 0.124 | | | P(T<=t) one-tail | 0.006 | | |
| t Critical one-tail | 2.015 | | | t Critical one-tail | 2.015 | | | t Critical one-tail | 2.015 | | |
| P(T<=t) two-tail | 0.830 | | | P(T<=t) two-tail | 0.249 | | | P(T<=t) two-tail | 0.012 | | |
| t Critical two-tail | 2.571 | | | t Critical two-tail | 2.571 | | | t Critical two-tail | 2.571 | | |
| Mercury | | 1997 | 1994 | Lead | | 1997 | 1994 | Lead | | 1997 | 1994 |
| Mean | 9.76 | 5.63 | | Mean | 57.4 | 76.2 | | Mean | 57.4 | 76.2 | |
| Variance | 59.677 | 37.237 | | Variance | 430.02 | 365.23 | | Variance | 430.02 | 365.23 | |
| Observations | 6 | 6 | | Observations | 6 | 6 | | Observations | 6 | 6 | |
| Pearson Correlation | 0.746 | | | Pearson Correlation | 0.816 | | | Pearson Correlation | 0.816 | | |
| Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | | Hypothesized Mean Difference | 0 | | |
| df | 5 | | | df | 5 | | | df | 5 | | |
| t Stat | 1.967 | | | t Stat | -3.792 | | | t Stat | -3.792 | | |
| P(T<=t) one-tail | 0.053 | | | P(T<=t) one-tail | 0.006 | | | P(T<=t) one-tail | 0.006 | | |
| t Critical one-tail | 2.015 | | | t Critical one-tail | 2.015 | | | t Critical one-tail | 2.015 | | |
| P(T<=t) two-tail | 0.106 | | | P(T<=t) two-tail | 0.012 | | | P(T<=t) two-tail | 0.013 | | |
| t Critical two-tail | 2.571 | | | t Critical two-tail | 2.571 | | | t Critical two-tail | 2.571 | | |

The change in concentration (delta) was also calculated on both particle size corrected and uncorrected data to compare concentrations between years at each station (Table 8). The most noteworthy changes include a decrease in concentrations of lead, phosphorus and % TOC at all six stations (with the exception of Pb at CS132). These decreases were statistically significant according to results from the paired t-test (Table 7). Nickel concentrations were lower in 1997 than 1994 at five stations but the decrease was not statistically significant. Additionally, Hg concentrations increased at four stations, decreased at station CS126 and generally remained the same at station CS156. However, these changes in concentration were also not statistically significant. A review of the QA/QC analysis (Appendix C2) indicates that the coefficient of variance for Hg within the box corer was high relative to the other parameters (range - 12% to 23%). This variability should be considered when comparing concentrations between years. As well, QA/QC data used to examine within station variability (Appendix C3) show a patchy distribution for total phosphorus making it difficult to distinguish changes in phosphorus concentrations over time from the inherent spatial variability in the sediment.

Any conclusions regarding changes in sediment quality between 1994 and 1997 for the study area based on this limited data set are tentative. It is unknown if three years between sampling dates is sufficient to measure real change in sediment contaminant concentrations in a river where sediment transport, deposition and resuspension on an annual basis may be significant.

Comparison of Surface Sample (3 cm) Concentrations With Concentrations in the Surface Core Samples (10 cm).

Since the sediment surface is subjected to constant mixing due to changes in flow, deposition of new sediment and resuspension of existing surface sediment in addition to mixing from bioturbation, differences in sediment concentration between the sediment layers requires cautious interpretation. An ANCOVA using particle size as the covariate was performed on the log transformed data to distinguish a difference between contaminant concentrations in surface samples and core samples. A test for heterogeneity of slopes showed that there was no significant difference between slopes for all parameters for the two sample types with the exception of lead ($p<0.02$) (Table 9). As a result lead was dropped from the analysis. Particle size accounted for a significant amount of the variation in sediment concentration for all parameters with the exception of zinc and total phosphorus.

The ANCOVA found that there was no significant difference in metal concentrations in the surface samples compared with the 10 cm core top samples for all the parameters tested with the exception of cadmium ($p<0.042$) and chromium ($p<0.036$) which showed weak statistical differences (Table 9). A Bonferroni correction to adjust the P value for multiple comparisons would likely find that these two parameters were not significantly different. This suggests that sediment quality has not changed at the surface relative to the underlying sediment for most parameters, however, sediment mixing due to bioturbation and other physical processes within the river may make any changes in sediment quality too small to measure.

TABLE 8: Surface sediment contaminant concentrations (ug/g) (10 cm core data) from stations sampled in 1994 and 1997 and the change (Delta) in concentration between years.
A negative Delta indicates that the concentration has increased from 1994 to 1997.

| Station number | Chromium | Delta Cr | Iron | 1994 | 1997 | Delta Fe | Mercury | 1994 | 1997 | Delta Hg | Manganese | 1994 | 1997 | Delta Mn |
|--------------------------|----------|----------|------|-------|-------|----------|---------|-------|--------|----------|-----------|------|------|----------|
| 1994 / 1997 | 1994 | 1997 | 1994 | 1997 | 1997 | 1994 | 1994 | 1994 | 1997 | 1997 | 1994 | 1994 | 1997 | 1997 |
| CS15 / CS115 | 46.7 | 41.8 | 4.9 | 19000 | 16400 | 2600 | 1.55 | 3.66 | -2.11 | 300 | 281 | 19 | | |
| CS26 / CS126 | 29.1 | 23.8 | 5.3 | 15061 | 12700 | 2361 | 10.94 | 5.20 | 5.74 | 257.98 | 201 | 57 | | |
| CS28 / CS128 | 35.1 | 32.7 | 2.4 | 17116 | 17900 | -784 | 3.86 | 11.20 | -7.34 | 321.47 | 345 | -24 | | |
| CS31 / CS131 | 36.0 | 33.3 | 2.7 | 16000 | 18300 | -2300 | 12.25 | 19.50 | -7.25 | 260.00 | 347 | -87 | | |
| CS32 / CS132 | 40.0 | 43.8 | -3.8 | 16269 | 17300 | -1031 | ND | 6.78 | -6.78 | 250.84 | 329 | -78 | | |
| CS56 / CS156 | 27.0 | 26.2 | 0.8 | 14000 | 14400 | -400 | 0.61 | 0.80 | -0.19 | 230.00 | 272 | -42 | | |
| CS31 / CS131 (3 cm grab) | 31.7 | 32.0 | -0.3 | 14528 | 14700 | -172 | 2.07 | 14.70 | -12.63 | 240.88 | 288 | -47 | | |

| Nickel | Delta Ni | Lead | Delta Pb | Phosphorus | Delta TP | %TOC | Delta TOC | | | | | | |
|--------------------------|----------|------|----------|------------|----------|------|-----------|-----|-----|------|------|-----|--|
| 1994 / 1997 | 1997 | 1994 | 1997 | 1994 | 1997 | 1994 | 1997 | | | | | | |
| CS15 / CS115 | 27.7 | 26.6 | 1.1 | 113.3 | 99.8 | 13.5 | 1170 | 678 | 492 | 4.01 | 2.45 | 1.6 | |
| CS26 / CS126 | 21.3 | 19.1 | 2.2 | 59.5 | 25.2 | 34.3 | 1170 | 457 | 713 | 5.00 | 3.07 | 1.9 | |
| CS28 / CS128 | 25.5 | 18.0 | 7.5 | 57.4 | 37.3 | 20.1 | 1400 | 681 | 719 | 5.00 | 2.85 | 2.2 | |
| CS31 / CS131 | 25.0 | 17.8 | 7.2 | 58.0 | 40.6 | 17.4 | 1360 | 659 | 701 | 4.64 | 4.04 | 0.6 | |
| CS32 / CS132 | 26.5 | 24.6 | 1.9 | 59.2 | 63.3 | -4.1 | 1150 | 818 | 332 | 4.02 | 2.52 | 1.5 | |
| CS56 / CS156 | 21.0 | 14.6 | 6.4 | 37.0 | 26.7 | 10.3 | 930 | 745 | 185 | 2.83 | 1.74 | 1.1 | |
| CS31 / CS131 (3 cm grab) | 20.8 | 24.9 | -4.1 | 40.2 | 33.7 | 6.5 | 1080 | 853 | 227 | 3.09 | 2.74 | 0.4 | |

Data Corrected for Particle Size - Sediment contaminant data were normalized to a fine (< 63 μ m) particle content of 74%.

| Station number | Chromium | Delta Cr | Iron | 1994 | 1997 | Delta Fe | Mercury | 1994 | 1997 | Delta Hg | Manganese | 1994 | 1997 | Delta Mn |
|--------------------------|----------|----------|-------|-------|-------|----------|---------|-------|--------|----------|-----------|------|------|----------|
| 1994 / 1997 | 1994 | 1997 | 1994 | 1997 | 1997 | 1994 | 1994 | 1994 | 1997 | 3.29 | -1.72 | 305 | 253 | 52 |
| CS15 / CS115 | 47.4 | 37.6 | 9.8 | 19286 | 14748 | 4538 | 1.57 | | | | | | | |
| CS26 / CS126 | 32.6 | 43.0 | -10.5 | 16828 | 22953 | -6125 | 12.22 | 9.40 | 2.83 | 288 | 363 | -75 | | |
| CS28 / CS128 | 42.1 | 41.0 | 1.1 | 20556 | 22443 | -1887 | 4.64 | 14.04 | -9.41 | 386 | 433 | -46 | | |
| CS31 / CS131 | 41.7 | 38.4 | 3.3 | 18539 | 21089 | -2550 | 14.19 | 22.47 | -8.28 | 301 | 400 | -99 | | |
| CS32 / CS132 | 48.5 | 53.2 | -4.7 | 19747 | 21015 | -1269 | ND | 8.24 | -8.24 | 304 | 400 | -95 | | |
| CS56 / CS156 | 49.5 | 37.8 | 11.7 | 25659 | 20760 | 4900 | 1.12 | 1.15 | -0.03 | 422 | 392 | 29 | | |
| CS31 / CS131 (3 cm grab) | 47.7 | 32.0 | 15.7 | 21858 | 14700 | 7158 | 3.11 | 14.70 | -11.59 | 362 | 288 | 74 | | |

| Nickel | Delta Ni | Lead | Delta Pb | Phosphorus | Delta TP | %TOC | Delta TOC | | | | | | |
|--------------------------|----------|------|----------|------------|----------|------|-----------|------|-----|------|------|------|--|
| 1994 / 1997 | 1997 | 1994 | 1997 | 1994 | 1997 | 1994 | 1997 | | | | | | |
| CS15 / CS115 | 28.1 | 23.9 | 4.2 | 115.0 | 89.7 | 25.3 | 1188 | 610 | 578 | 4.07 | 2.20 | 1.87 | |
| CS26 / CS126 | 23.8 | 34.5 | -10.7 | 66.5 | 45.5 | 21.0 | 1307 | 826 | 481 | 5.59 | 5.55 | 0.04 | |
| CS28 / CS128 | 30.6 | 22.6 | 8.0 | 68.9 | 46.8 | 22.1 | 1681 | 854 | 828 | 6.00 | 3.57 | 2.43 | |
| CS31 / CS131 | 29.0 | 20.5 | 8.5 | 67.2 | 46.8 | 20.4 | 1576 | 759 | 816 | 5.38 | 4.65 | 0.72 | |
| CS32 / CS132 | 32.2 | 29.9 | 2.3 | 71.8 | 76.9 | -5.1 | 1396 | 994 | 402 | 4.88 | 3.06 | 1.82 | |
| CS56 / CS156 | 38.5 | 21.0 | 17.4 | 67.8 | 38.5 | 29.3 | 1705 | 1074 | 630 | 5.19 | 2.51 | 2.68 | |
| CS31 / CS131 (3 cm grab) | 31.3 | 24.9 | 6.4 | 60.5 | 33.7 | 26.8 | 1625 | 853 | 772 | 4.65 | 2.74 | 1.91 | |

ND - not detected

TABLE 9: Results of the analysis of covariance for each term included in the general linear model testing for sample type (10 cm core samples vs 3 cm surface samples) variability. The data is reported for each parameter measured in sediment from the St. Lawrence River, 1997.

| Contaminant | | F _* ratio | P | Contaminant | | F _* ratio | P |
|-------------|----------------------------|----------------------|--------|----------------------|----------------------------|----------------------|--------|
| Aluminum | Particle size | 64.57 | 0.0001 | Manganese | Particle size | 16.91 | 0.0003 |
| | Sample type | 1.39 | 0.249 | | Sample type | 0.11 | 0.7479 |
| | Particle size *sample type | 0.68 | 0.4169 | | Particle size *sample type | 0.75 | 0.3953 |
| | | | | | | | |
| Cadmium | Particle size | 50.95 | 0.0001 | Nickel | Particle size | 81.46 | 0.0001 |
| | Sample type | 4.58 | 0.0418 | | Sample type | 0.01 | 0.9307 |
| | Particle size *sample type | 0 | 0.9703 | | Particle size *sample type | 0.51 | 0.4804 |
| | | | | | | | |
| Chromium | Particle size | 72.88 | 0.0001 | Lead | Particle size | 20.25 | 0.0001 |
| | Sample type | 4.87 | 0.0364 | | Sample type | 1.81 | 0.1896 |
| | Particle size *sample type | 0.75 | 0.3934 | | Particle size *sample type | 6.76 | 0.0152 |
| | | | | | | | |
| Copper | Particle size | 13.53 | 0.0011 | Zinc | Particle size | 1.22 | 0.2799 |
| | Sample type | 3.96 | 0.0572 | | Sample type | 1.86 | 0.1848 |
| | Particle size *sample type | 0.73 | 0.401 | | Particle size *sample type | 0.5 | 0.4876 |
| | | | | | | | |
| Iron | Particle size | 30.08 | 0.0001 | Total Phosphorus | Particle size | 1.87 | 0.1831 |
| | Sample type | 0.42 | 0.5208 | | Sample type | 1.09 | 0.3063 |
| | Particle size *sample type | 2.1 | 0.1591 | | Particle size *sample type | 0.6 | 0.4446 |
| | | | | | | | |
| Mercury | Particle size | 5.5 | 0.0269 | Total Organic Carbon | Particle size | 4.41 | 0.0456 |
| | Sample type | 1.14 | 0.2954 | | Sample type | 0.12 | 0.7357 |
| | Particle size *sample type | 1.91 | 0.1784 | | Particle size *sample type | 2.92 | 0.0991 |
| | | | | | | | |

CONCLUSIONS

The following conclusions have been based on the stated survey objectives.

1) *Surface sediment located in the deposition zone about 1.4 km downstream of the Domtar/ICI diffuser has mercury concentrations greater than the LEL but less than the SEL.*

Sediment at this site was historically contaminated with mercury. This is evident from high concentrations in sediment collected from the bottom of the core (4.63 ug/g and 13.70 ug/g) relative to mercury concentrations in sediment at the top of the core sample (1.19 ug/g and 1.71 ug/g respectively).

This area was not contaminated with other metals.

2) *Sediment located adjacent to the north east side of Cornwall Island was not contaminated with metals.*

3) *The highest concentrations of mercury, lead, copper and zinc in the study area were located downstream of the former Courtaulds Fibre Canada facility. Remediation of contaminated sediment, if considered, should be focussed in this deposition zone. Sampling sites in this area could be used for future monitoring to assess temporal changes in sediment quality.*

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APPENDIX

APPENDIX A: Station location for sediment sampling survey in the St. Lawrence River, 1997.

| Station | Sampling Date | Northing | Easting |
|---------|---------------|------------|----------|
| CS105 | 97-10-22 | 4984819.60 | 523938.0 |
| CS109 | 97-10-22 | 4984878.70 | 523978.8 |
| CS115 | 97-10-22 | 4984932.70 | 524087.5 |
| CS117 | 97-10-22 | 4985001.00 | 524182.0 |
| CS126 | 97-10-23 | 4985038.20 | 524381.7 |
| CS127 | 97-10-22 | 4985064.40 | 524391.8 |
| CS128 | 97-10-23 | 4985063.90 | 524479.0 |
| CS131 | 97-10-23 | 4985169.20 | 524567.2 |
| CS132 | 97-10-23 | 4985195.00 | 524621.7 |
| CS135 | 97-10-23 | 4985277.50 | 524738.5 |
| CS156 | 97-10-21 | 4985718.00 | 525505.1 |
| CS164 | 97-10-22 | 4984943.30 | 524067.5 |
| CS166 | 97-10-21 | 4984027.00 | 521105.6 |
| CS167 | 97-10-22 | 4984039.00 | 521149.9 |
| CS168 | 97-10-21 | 4984060.60 | 521183.4 |
| CS171 | 97-10-21 | 4985668.00 | 526865.0 |
| CS172 | 97-10-21 | 4985955.70 | 527025.8 |
| CS173 | 97-10-23 | 4984838.00 | 525372.6 |
| CS175 | 97-10-23 | 4984852.30 | 525541.7 |
| CS176 | 97-10-23 | 4984773.90 | 525632.6 |
| CS177 | 97-10-23 | 4984825.00 | 525763.0 |
| CS179 | 97-10-23 | 4984798.60 | 525940.0 |
| CS181 | 97-10-21 | 4984785.00 | 526195.0 |
| CS182 | 97-10-21 | 4984827.90 | 526295.7 |

| 1994 station locations | Northing | Easting |
|------------------------|-----------|----------|
| CS05 | 4984856.3 | 523925.8 |
| CS09 | 4984887.7 | 523966.0 |
| CS15 | 4984933.8 | 524079.1 |
| CS17 | 4984991.1 | 524180.5 |
| CS26 | 4985031.2 | 524382.7 |
| CS27 | 4985076.7 | 524403.7 |
| CS28 | 4985061.0 | 524478.3 |
| CS31 | 4985167.6 | 524564.7 |
| CS32 | 4985190.8 | 524620.1 |
| CS35 | 4985276.6 | 524722.7 |
| CS56 | 4985722.6 | 525503.0 |
| CS64 | 4984953.3 | 524056.6 |
| CS31-S | 4985172.5 | 524564.1 |

QA/QC site locations

| | | | |
|-------|----------|------------------|-----------------------|
| CS126 | 97-10-23 | 4985031.5 | 524387.2 |
| CS164 | 97-10-22 | 4984945.7 | 524067.9 |
| CS167 | 97-10-22 | 4984040.5 | 521143.4 |
| CS176 | 97-10-23 | 4984776.7 | 525619.3 |
| CS135 | 97-10-23 | Data unavailable | Samples not submitted |

Estimated Core Lengths

| | |
|-------|-------|
| CS166 | 29 cm |
| CS167 | 20 cm |
| CS168 | 23 cm |
| CS175 | 24 cm |
| CS179 | 20 cm |
| CS181 | 24 cm |

APPENDIX B(1): Particle size correction. Sediment contaminant data (10 cm surface core samples) were normalized to aluminum concentrations.
 St. Lawrence River, 1997. B - bottom 10 cm of the core. S - surface sample (top 3 cm).

| Station Number | Cd:Al | Cr:Al | Cu:Al | Fe:Al | Hg:Al | Mn:Al | Ni:Al | P:Al | Pb:Al | Zn:Al | TOC:Al |
|----------------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|
| CS105 | 0.000049 | 0.002388 | 0.002301 | 0.984694 | 0.000085 | 0.015051 | 0.001367 | 0.038827 | 0.003230 | 0.030612 | 1.492857 |
| CS109 | 0.000071 | 0.002667 | 0.003358 | 1.006061 | 0.000293 | 0.015273 | 0.001588 | 0.038667 | 0.008242 | 0.046000 | 1.740606 |
| CS115 | 0.000094 | 0.002714 | 0.003760 | 1.064935 | 0.000238 | 0.018247 | 0.001727 | 0.044026 | 0.006481 | 0.039156 | 1.588312 |
| CS117 | 0.000053 | 0.002730 | 0.003035 | 1.156522 | 0.000174 | 0.021304 | 0.001730 | 0.047739 | 0.007261 | 0.024261 | 2.046087 |
| CS126 | 0.000075 | 0.002471 | 0.002762 | 1.318795 | 0.000540 | 0.020872 | 0.001983 | 0.047456 | 0.002617 | 0.012980 | 3.194185 |
| CS127 | 0.000035 | 0.002059 | 0.002178 | 1.007407 | 0.000320 | 0.020741 | 0.001022 | 0.040593 | 0.003970 | 0.011259 | 1.245185 |
| CS128 | 0.000038 | 0.001946 | 0.002000 | 1.065476 | 0.000667 | 0.020536 | 0.001071 | 0.040536 | 0.002220 | 0.010536 | 1.696429 |
| CS131 | 0.000038 | 0.001959 | 0.002500 | 1.076471 | 0.001147 | 0.020412 | 0.001047 | 0.038765 | 0.002388 | 0.010941 | 2.374706 |
| CS132 | 0.000044 | 0.002009 | 0.002794 | 0.793578 | 0.000311 | 0.015092 | 0.001128 | 0.037523 | 0.002904 | 0.021743 | 1.154128 |
| CS135 | 0.000052 | 0.002121 | 0.002395 | 0.973684 | 0.000197 | 0.017421 | 0.001205 | 0.034737 | 0.003095 | 0.012316 | 1.109474 |
| CS156 | 0.000043 | 0.002298 | 0.001482 | 1.263158 | 0.000070 | 0.023860 | 0.001281 | 0.065351 | 0.002342 | 0.011754 | 1.526316 |
| CS164 | 0.000056 | 0.003044 | 0.003744 | 1.106250 | 0.000191 | 0.018250 | 0.001556 | 0.052938 | 0.008500 | 0.041813 | 1.533750 |
| CS166 | 0.000050 | 0.001947 | 0.001770 | 0.942623 | 0.000033 | 0.014754 | 0.001127 | 0.047131 | 0.001475 | 0.005410 | 1.326230 |
| CS167 | 0.000054 | 0.002202 | 0.002363 | 0.970238 | 0.000071 | 0.015774 | 0.001780 | 0.051607 | 0.002179 | 0.006369 | 1.767857 |
| CS168 | 0.000052 | 0.002206 | 0.001865 | 1.000000 | 0.000101 | 0.017059 | 0.001412 | 0.055706 | 0.002235 | 0.006000 | 1.066471 |
| CS171 | 0.000037 | 0.001962 | 0.001683 | 1.026882 | 0.000023 | 0.019301 | 0.001108 | 0.049355 | 0.001591 | 0.006344 | 1.083871 |
| CS172 | 0.000036 | 0.001885 | 0.001478 | 0.852564 | 0.000020 | 0.014071 | 0.001093 | 0.035897 | 0.001410 | 0.011218 | 1.070513 |
| CS173 | 0.000037 | 0.001786 | 0.001410 | 0.883792 | 0.000004 | 0.014954 | 0.001064 | 0.035168 | 0.001202 | 0.004740 | 0.957187 |
| CS175 | 0.000035 | 0.001795 | 0.001360 | 0.907534 | 0.000003 | 0.015890 | 0.001041 | 0.036986 | 0.001284 | 0.004658 | 0.867123 |
| CS176 | 0.000035 | 0.001775 | 0.001390 | 0.900901 | 0.000005 | 0.017778 | 0.001063 | 0.035435 | 0.001171 | 0.004685 | 0.930631 |
| CS177 | 0.000034 | 0.001731 | 0.001332 | 0.909091 | 0.000004 | 0.017308 | 0.001077 | 0.035664 | 0.001147 | 0.004615 | 0.908741 |
| CS179 | 0.000036 | 0.001794 | 0.001379 | 0.939716 | 0.000005 | 0.017305 | 0.001117 | 0.038652 | 0.001248 | 0.004823 | 0.789716 |
| CS181 | 0.000040 | 0.001878 | 0.001396 | 0.952941 | 0.000005 | 0.017059 | 0.001173 | 0.039059 | 0.001302 | 0.004941 | 0.796078 |
| CS182 | 0.000037 | 0.001802 | 0.001309 | 0.967078 | 0.000005 | 0.020247 | 0.001169 | 0.044856 | 0.001119 | 0.004815 | 1.031687 |
| CS-166-B | 0.000048 | 0.001967 | 0.001707 | 0.870370 | 0.000005 | 0.012148 | 0.001215 | 0.042222 | 0.001637 | 0.005259 | 1.208519 |
| CS-167-B | 0.000057 | 0.001781 | 0.002607 | 0.607287 | 0.000187 | 0.010000 | 0.001571 | 0.045344 | 0.002360 | 0.006397 | 2.730769 |
| CS-168-B | 0.000062 | 0.002012 | 0.002934 | 0.664063 | 0.000535 | 0.011172 | 0.001402 | 0.042578 | 0.002875 | 0.007305 | 2.272656 |
| CS-175-B | 0.000060 | 0.002109 | 0.001583 | 0.939130 | 0.000011 | 0.015609 | 0.001339 | 0.040478 | 0.002035 | 0.005696 | 0.976522 |
| CS-179-B | 0.000052 | 0.002034 | 0.001471 | 0.995098 | 0.000008 | 0.016863 | 0.001319 | 0.044559 | 0.001775 | 0.005294 | 0.953431 |

| Station Number | Cd:Al | Cr:Al | Cu:Al | Fe:Al | Hg:Al | Mn:Al | Ni:Al | P:Al | Pb:Al | Zn:Al | TOC:Al |
|----------------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|----------|
| CS-181-B | 0.000056 | 0.002083 | 0.001568 | 0.973799 | 0.00009 | 0.015983 | 0.001362 | 0.041572 | 0.001891 | 0.005415 | 0.934934 |
| CS-109-S | 0.000049 | 0.002584 | 0.002926 | 0.978947 | 0.000305 | 0.016684 | 0.001374 | 0.053684 | 0.008211 | 0.035421 | 1.448947 |
| CS-115-S | 0.000056 | 0.002253 | 0.002969 | 1.00000 | 0.000084 | 0.017474 | 0.001443 | 0.050361 | 0.003119 | 0.017268 | 1.168041 |
| CS-117-S | 0.000046 | 0.002171 | 0.002289 | 0.994652 | 0.000072 | 0.017807 | 0.001380 | 0.051604 | 0.003251 | 0.014439 | 1.095722 |
| CS-126-S | 0.000035 | 0.002336 | 0.001787 | 1.540984 | 0.000098 | 0.029262 | 0.001311 | 0.069836 | 0.001615 | 0.006721 | 2.442623 |
| CS-128-S | 0.000035 | 0.002437 | 0.002049 | 1.475728 | 0.000334 | 0.028447 | 0.001408 | 0.074272 | 0.002291 | 0.006699 | 1.845631 |
| CS-131-S | 0.000051 | 0.002148 | 0.002523 | 0.986577 | 0.000987 | 0.019329 | 0.001671 | 0.057248 | 0.002262 | 0.016846 | 1.836913 |
| CS-135-S | 0.000051 | 0.002211 | 0.001569 | 1.162602 | 0.000099 | 0.023415 | 0.001715 | 0.061626 | 0.002211 | 0.009512 | 1.425203 |
| CS-164-S | 0.000054 | 0.002595 | 0.003196 | 1.074324 | 0.000134 | 0.019662 | 0.001703 | 0.056824 | 0.006757 | 0.026689 | 1.906757 |
| CS-166-S | 0.000048 | 0.001936 | 0.001855 | 0.948718 | 0.000045 | 0.014316 | 0.001291 | 0.042137 | 0.001440 | 0.005684 | 1.491880 |
| CS-167-S | 0.000056 | 0.002189 | 0.001934 | 1.122951 | 0.000141 | 0.019590 | 0.001598 | 0.053525 | 0.001795 | 0.006303 | 1.942623 |
| CS-168-S | 0.000064 | 0.002271 | 0.001797 | 1.110169 | 0.000059 | 0.018814 | 0.001763 | 0.061356 | 0.002093 | 0.006839 | 1.845763 |
| CS-176-S | 0.000040 | 0.001771 | 0.001509 | 0.952727 | 0.000005 | 0.024182 | 0.001396 | 0.040364 | 0.001073 | 0.004945 | 1.174182 |
| CS-179-S | 0.000044 | 0.001795 | 0.001423 | 0.958159 | 0.000005 | 0.018619 | 0.001410 | 0.040000 | 0.001109 | 0.005021 | 1.288285 |
| CS-181-S | 0.000042 | 0.001829 | 0.001410 | 0.958525 | 0.000006 | 0.017788 | 0.001309 | 0.032166 | 0.001244 | 0.005069 | 1.164977 |
| CS-182-S | 0.000040 | 0.001804 | 0.001421 | 0.978723 | 0.000005 | 0.021830 | 0.001374 | 0.041957 | 0.001119 | 0.004936 | 1.202979 |

APPENDIX B(2): Particle size correction. Sediment contaminant data (10 cm surface core samples) were normalized to a fine (< 63 μ m) particle content of 74%. St. Lawrence River, 1997. B - bottom 10 cm of the core. S - surface sample (top 3 cm).

| Station | Al Corrected mg/Kg | Cd Corrected mg/Kg | Cr Corrected mg/Kg | Cu Corrected mg/Kg | Fe Corrected mg/Kg | Hg Corrected mg/Kg | Mn Corrected mg/Kg | Ni Corrected mg/Kg | Pb Corrected mg/Kg | TP Corrected mg/Kg | TOC Corrected % |
|---------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|--------------------|-----------------|
| CS105 | 17228 | 0.837 | 41.1 | 39.6 | 16965 | 1.47 | 259 | 23.6 | 55.6 | 527 | 669 |
| CS109 | 14250 | 1.010 | 38.0 | 47.8 | 14336 | 4.17 | 218 | 22.6 | 117.5 | 656 | 551 |
| CS115 | 13849 | 1.304 | 37.6 | 52.1 | 14748 | 3.29 | 253 | 23.9 | 89.7 | 542 | 610 |
| CS117 | 17462 | 0.923 | 47.7 | 53.0 | 20195 | 3.04 | 372 | 30.2 | 126.8 | 424 | 834 |
| CS126 | 17404 | 1.301 | 43.0 | 48.1 | 22953 | 9.40 | 363 | 34.5 | 45.5 | 226 | 826 |
| CS127 | 31858 | 1.116 | 65.6 | 69.4 | 32094 | 10.19 | 661 | 32.6 | 126.5 | 359 | 1293 |
| CS128 | 21064 | 0.799 | 41.0 | 42.1 | 22443 | 14.04 | 433 | 22.6 | 46.8 | 222 | 854 |
| CS131 | 19591 | 0.743 | 38.4 | 49.0 | 21089 | 22.47 | 400 | 20.5 | 46.8 | 214 | 759 |
| CS132 | 26482 | 1.153 | 53.2 | 74.0 | 21015 | 8.24 | 400 | 29.9 | 76.9 | 576 | 994 |
| CS135 | 21076 | 1.092 | 44.7 | 50.5 | 20522 | 4.16 | 367 | 25.4 | 65.2 | 260 | 732 |
| CS156 | 16435 | 0.706 | 37.8 | 24.4 | 20760 | 1.15 | 392 | 21.0 | 38.5 | 193 | 1074 |
| CS164 | 16699 | 0.927 | 50.8 | 62.5 | 18473 | 3.19 | 305 | 26.0 | 141.9 | 698 | 884 |
| CS166 | 19568 | 0.978 | 38.1 | 34.6 | 18445 | 0.64 | 289 | 22.1 | 28.9 | 106 | 922 |
| CS167 | 17266 | 0.935 | 38.0 | 40.8 | 16752 | 1.22 | 272 | 30.7 | 37.6 | 110 | 891 |
| CS168 | 21719 | 1.120 | 47.9 | 40.5 | 21719 | 2.18 | 371 | 30.7 | 48.5 | 130 | 1210 |
| CS171 | 21302 | 0.790 | 41.8 | 35.8 | 21875 | 0.50 | 411 | 23.6 | 33.9 | 135 | 1051 |
| CS172 | 25688 | 0.922 | 48.4 | 38.0 | 21901 | 0.51 | 361 | 28.1 | 36.2 | 288 | 922 |
| CS173 | 25614 | 0.948 | 45.7 | 36.1 | 22637 | 0.10 | 383 | 27.3 | 30.8 | 121 | 901 |
| CS175 | 24842 | 0.868 | 44.6 | 33.8 | 22545 | 0.07 | 395 | 25.9 | 31.9 | 116 | 919 |
| CS176 | 26306 | 0.908 | 46.7 | 36.6 | 23699 | 0.12 | 468 | 28.0 | 30.8 | 123 | 932 |
| CS177 | 24472 | 0.835 | 42.4 | 32.6 | 22247 | 0.10 | 424 | 26.4 | 28.1 | 113 | 873 |
| CS179 | 24355 | 0.881 | 43.7 | 33.6 | 22886 | 0.12 | 421 | 27.2 | 30.4 | 117 | 941 |
| CS181 | 21351 | 0.846 | 40.1 | 29.8 | 20346 | 0.11 | 364 | 25.0 | 27.8 | 105 | 834 |
| CS182 | 20938 | 0.776 | 37.7 | 27.4 | 20248 | 0.10 | 424 | 24.5 | 23.4 | 101 | 939 |
| CS199-S | 15471 | 0.756 | 40.0 | 45.3 | 15146 | 4.71 | 258 | 21.3 | 127.0 | 548 | 831 |
| CS115-S | 21488 | 1.196 | 48.4 | 63.8 | 21488 | 1.81 | 375 | 31.0 | 67.0 | 371 | 1082 |
| CS117-S | 24015 | 1.116 | 52.1 | 55.0 | 23887 | 1.73 | 428 | 33.1 | 78.1 | 347 | 1239 |
| CS126-S | 20361 | 0.708 | 47.6 | 36.4 | 31376 | 2.00 | 596 | 26.7 | 32.9 | 137 | 1422 |
| CS128-S | 22317 | 0.778 | 54.4 | 45.7 | 32933 | 7.45 | 635 | 31.4 | 51.1 | 150 | 1658 |
| CS131-S | 14900 | 0.758 | 32.0 | 37.6 | 14700 | 14.70 | 288 | 24.9 | 33.7 | 251 | 853 |
| CS135-S | 14868 | 0.754 | 32.9 | 23.3 | 17286 | 1.47 | 348 | 25.5 | 32.9 | 141 | 916 |
| CS164-S | 13134 | 0.706 | 34.1 | 42.0 | 14110 | 1.76 | 258 | 22.4 | 88.7 | 351 | 746 |
| CS166-S | 19117 | 0.915 | 37.0 | 35.5 | 18137 | 0.87 | 274 | 24.7 | 27.5 | 109 | 806 |
| CS167-S | 14282 | 0.804 | 31.3 | 27.6 | 16037 | 2.01 | 280 | 22.8 | 25.6 | 90 | 764 |

| Station | Al Corrected mg/Kg | Cd Corrected mg/Kg | Cr Corrected mg/Kg | Cu Corrected mg/Kg | Fe Corrected mg/Kg | Hg Corrected mg/Kg | Mn Corrected mg/Kg | Ni Corrected mg/Kg | Pb Corrected mg/Kg | Zn Corrected mg/Kg | TP Corrected mg/Kg | TOC Corrected % |
|---------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|--------------------|
| CS168-S | 15927 | 1.018 | 36.2 | 28.6 | 17681 | 0.94 | 300 | 28.1 | 33.3 | 109 | 977 | 2.94 |
| CS176-S | 21138 | 0.846 | 37.4 | 31.9 | 20139 | 0.12 | 511 | 29.5 | 22.7 | 105 | 863 | 2.48 |
| CS179-S | 19989 | 0.870 | 35.9 | 28.4 | 19152 | 0.10 | 372 | 28.2 | 22.2 | 100 | 800 | 2.58 |
| CS181-S | 18231 | 0.770 | 33.4 | 25.7 | 17475 | 0.11 | 324 | 23.9 | 22.7 | 92 | 586 | 2.12 |
| CS182-S | 19305 | 0.771 | 34.8 | 27.4 | 18895 | 0.10 | 421 | 26.5 | 21.6 | 95 | 810 | 2.32 |
| CS186-B | 21307 | 1.018 | 41.9 | 36.4 | 18545 | 0.10 | 259 | 25.9 | 34.9 | 112 | 900 | 2.57 |
| CS167-B | 20069 | 1.154 | 35.8 | 52.3 | 12188 | 3.76 | 201 | 31.5 | 47.4 | 128 | 910 | 5.48 |
| CS168-B | 19678 | 1.222 | 39.6 | 57.7 | 13067 | 10.53 | 220 | 27.6 | 56.6 | 144 | 838 | 4.47 |
| CS175-B | 19749 | 1.185 | 41.6 | 31.3 | 18546 | 0.21 | 308 | 26.4 | 40.2 | 112 | 799 | 1.93 |
| CS179-B | 19405 | 1.018 | 39.5 | 28.5 | 19310 | 0.16 | 327 | 25.6 | 34.4 | 103 | 865 | 1.85 |
| CS181-B | 20249 | 1.141 | 42.2 | 31.7 | 19719 | 0.18 | 324 | 27.6 | 38.3 | 110 | 842 | 1.89 |

APPENDIX B(3I): Particle size correction. Sediment contaminant data were normalized to aluminum concentrations. Sediment samples were collected from the St. Lawrence River in 1991 at Maitland, Ontario. The mean, standard deviation and 95% upper and lower confidence intervals (CI) were generated from all the survey data from the study area. Maitland (n=116).

| Maitland Data | Cd Al | Cr Al | Cu Al | Fe Al | Hg Al | Mn Al | Pb Al | Zn Al | TP Al | TOC Al |
|--------------------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|-----------|
| Mean | 7.842E-05 | 2.765E-03 | 2.484E-03 | 1.914E-00 | 7.485E-06 | 4.947E-02 | 1.977E-03 | 2.883E-03 | 7.308E-03 | 1.335E-04 |
| Standard Deviation | 5.567E-05 | 4.651E-04 | 1.264E-03 | 7.845E-01 | 6.449E-06 | 2.029E-02 | 3.647E-04 | 3.619E-03 | 2.401E-03 | 6.656E-05 |
| 95% upper CI | 8.855E-05 | 2.849E-03 | 2.714E-03 | 2.057E+00 | 8.658E-06 | 5.316E-02 | 2.043E-03 | 3.542E-03 | 7.744E-03 | 1.456E-04 |
| 95% lower CI | 6.829E-05 | 2.680E-03 | 2.254E-03 | 1.772E+00 | 6.311E-06 | 4.578E-02 | 1.911E-03 | 2.225E-03 | 6.871E-03 | 1.214E-04 |

APPENDIX B(3II): Particle size correction. Sediment contaminant data were normalized to a fine (< 63 μ m) particle content of 74%. Sediment samples collected from the St. Lawrence River in 1991 at Maitland, Ontario. The mean, standard deviation and 95% upper and lower confidence intervals (CI) were generated from all the survey data from the study area (ug/g dry weight). Maitland (n=116).

| Maitland Data | Cd Corrected | Cr Corrected | Cu Corrected | Fe Corrected | Hg Corrected | Mn Corrected | Ni Corrected | Pb Corrected | Zn Corrected | TP Corrected | TOC Corrected (%) | Al Corrected |
|--------------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|--------------|-------------------|--------------|
| Mean | 2.17 | 85.67 | 67.01 | 62.88 | 0.23 | 1562 | 59 | 83 | 211 | 4742 | 4.8 | 30906 |
| Standard Deviation | 2.12 | 64.16 | 36.60 | 52278 | 0.32 | 1135 | 34 | 122 | 124 | 6210 | 6.3 | 22222 |
| 95% upper CI | 2.55 | 97.35 | 73.67 | 71701 | 0.29 | 1768 | 65 | 106 | 233 | 5873 | 5.9 | 34950 |
| 95% lower CI | 1.78 | 73.99 | 60.35 | 52674 | 0.17 | 1355 | 52 | 61 | 188 | 3612 | 3.6 | 26862 |

Data from Richman and Townsend 1997

APPENDIX C

Quality Analysis\Quality Control - Among and Within-Station Variability.

Variability in sediment contaminant concentrations for all parameters among and within stations was compared using field replicates for core samples collected from four stations (CS126, CS135, CS164, CS167). Three replicate core samples were collected from within a single mini-box corer at each of the four stations. These results were used to assess among station variability and the variability associated within the mini-box corer. At each of the QA/QC stations a routine sample was also collected for the analyses described throughout this report. This sample which was collected from a separate drop of the mini-box corer was compared with the QA/QC samples from the same station to assess within station variability. Within sample variability was assessed by collecting split samples from one station (CS135). The three way split of each of two core samples from station CS135 was used to assess the variability associated with sample homogenization, possible sample contamination from handling practices, and laboratory “analytical error”. Data for within sample variability was available for all parameters with the exception of mercury. All raw data are provided in Appendix C4.

An analysis of variance (ANOVA) was performed on both log transformed and non-transformed core data to estimate the proportion of the total variability associated with the between station and within mini-box core components (Appendix C1). The percent variability associated with each category was calculated by first calculating the variance components using the expected mean squares for each category and then comparing the variance components with the total variance.

Using the non-transformed data, the lowest source of variability in the core samples was associated with the within mini-box component for all parameters (with the exception of iron, manganese and total phosphorus). The percent variability associated with the two components were similar for chromium. This suggests that sediment within the confines of the mini-box corer was homogeneous for most parameters. The ANOVA found a significant difference between station concentrations (p values ranged from 0.0001 to 0.04) for all parameters tested with the exception iron, manganese and total phosphorus. Accordingly, for all parameters with the exception of iron, manganese and total phosphorus, the total variability within the mini-box corer was low enough that differences between stations could be determined.

The results of the ANOVA suggest that for iron, manganese and total phosphorus (chromium to a lesser extent), the variability in sediment concentrations within the box was greater than between stations. However, these results were likely due to the narrow range in concentrations for these parameters throughout the study area rather than due to large differences within the mini-box corer. Differences in concentrations between stations were not identified because the concentrations of these parameters in the sediment were similar among the stations.

The ANOVA using log transformed data provided the same results as described above.

For most parameters, variability (expressed as the coefficient of variance - CV Appendix C2) within the mini-box corer was low (less than 12%) based on the replicate core data, with only a few exceptions. One station (CS126) had consistently higher CV's for most parameters than at the other three stations although they were still low and ranged from 4% to 23%. The CVs for all parameters at station CS164, CS167 and CS135 were less than 12% with only one exception at station CS164 where the CV for lead was 23% and at station CS167 where the CV for mercury was 18%. This suggests that variability in sediment quality within the mini-box corer was low but did vary between stations and with the parameter. Mercury and lead had the highest within box core variability.

Unlike the other QA/QC stations, samples collected from station C135 were a combination of one single sample and two additional samples that were each split into three sub-samples. Hence the CV for each parameters was calculated by two different methods. The CV was determined based on a mean that was calculated by using the first sample from each split sample and the third replicate (which was a single sample) as well as by using the mean from each split sample combined with the third replicate single sample. In both cases the CV for all parameters were almost identical, and with only one exception all CVs were less than 10%. These results, overall, suggest that the mini-box corer was a useful tool to accommodate a sampling design that required several samples from a single site since variability within the box was low. As well, the results suggest that for most parameters individual samples collected from the mini-box corer adequately described the sampling device without replication.

Variability for almost all parameters was low within split samples (station CS135) suggesting that the samples were well mixed prior to distribution among sample jars (Appendix C2). Based on laboratory QA/QC information some variability between split samples may be due to analytical error. The CVs for the two split samples for all parameters ranged from 1 to 21% with values for most parameters less than 10%. The highest variability was associated with total phosphorus and lead for both split samples and with zinc (14%) for one split while the other split had a CV for zinc of only 5%.

Within station variability was assessed by comparing the concentration of the parameter in the routine sample core with the mean and 95% upper and lower confidence intervals generated for the three replicate samples from the QA/QC mini-box core (Appendix C3i & C3ii). Comparisons were made using both particle size corrected and uncorrected data. Results were similar using both data sets since particle size differences between samples at any one station were low. Because the sample size was small (n=3), a t-test would have produced results with a predisposition to accepting the null hypothesis since the power of the test would have been low.

Results from the comparison between the routine stations with the QA/QC samples suggest some variability of contaminant concentrations within a station (particularly for total phosphorus and to a lesser extent for lead), but typically the difference in concentrations were not large. For most parameters the concentrations for the routine samples tended to be within the upper and lower confidence bands generated for the QA/QC samples. Samples which fell outside the confidence bands were not sufficiently different to be environmentally significant (i.e. the difference in concentration in almost all cases did not change the designation of whether the sample was above

of below the sediment quality guidelines). Variability within these stations appears to be low suggesting that single samples adequately represent a site. A review of the station's northing and easting for the QA/QC samples (Appendix A) show that QA/QC stations CS164 and CS167 were fairly close to the original samples.

For station CS126 the QA/QC station location difference was about 5.5 m east and 7 m south of the original site, possibly providing an explanation for the higher variability in contaminant concentrations observed at that station relative to the other QA/QC sites. The confidence bands generated for this station were also wide due to the high variability within the mini-box. Relatively high variability within the box and within the site suggest that this area may have a patchy sediment quality.

For QA/QC station CS135 mercury concentrations were only provided for the single replicate samples. However, the samples collected for the EC/MOE interlaboratory comparison were also collected from the same mini-box as the QA/QC samples which, therefore, essentially provided an additional replicate for the QA/QC mini-box drop. The mercury concentration in the single replicate was 11.3 $\mu\text{g/g}$ while the concentration in the interlab comparison sample was 9.41 $\mu\text{g/g}$. Both concentrations were high compared with the mercury concentration in the routine sample (3.75 $\mu\text{g/g}$). Information on the exact location of the QA/QC sample was not available to determine the distance between each drop of the mini-box corer. Speculation on within site variability for Hg at this station is therefore not possible.

In general, for the four QA/QC stations, there was good agreement for all parameters between the EC samples from the interlaboratory comparison with the three replicates collected for QA/QC analysis (Appendix E).

APPENDIX C(1): The expected mean squares generated by the ANOVA using non-transformed data were used to calculate the variance components associated with the between station variability and within station variability. The percent 'variability' within each component was calculated relative to the "total" variability based on the total variance component.

| | Variance Component | Percent Variability | | Variance Component | Percent Variability |
|-----------|--------------------|---------------------|------------------|--------------------|---------------------|
| Aluminum | | | Nickel | | |
| Total | 3628438.907 | | Total | 12.740 | |
| STN | 2870641.657 | 79 | STN | 7.885 | 62 |
| REP(STN) | 757797.250 | 21 | REP(STN) | 4.855 | 38 |
| Cadmium | | | Lead | | |
| Total | 0.048 | | Total | 4387.637 | |
| STN | 0.043 | 91 | STN | 4005.757 | 91 |
| REP(STN) | 0.004 | 9 | REP(STN) | 381.880 | 9 |
| Chromium | | | Zinc | | |
| Total | 29.895 | | Total | 46366.380 | |
| STN | 15.750 | 53 | STN | 45650.630 | 98 |
| REP(STN) | 14.144 | 47 | REP(STN) | 715.750 | 2 |
| Copper | | | Total Phosphorus | | |
| Total | 174.146 | | Total | 2836.361 | |
| STN | 163.586 | 94 | STN | 856.528 | 30 |
| REP(STN) | 10.559 | 6 | REP(STN) | 1979.833 | 70 |
| Iron | | | TOC | | |
| Total | 766211.130 | | Total | 1.200 | |
| STN | 375088.880 | 49 | STN | 1.171 | 98 |
| REP(STN) | 391122.250 | 51 | REP(STN) | 0.028 | 2 |
| Manganese | | | Mercury | | |
| Total | 293.565 | | Total | 17.128 | |
| STN | 60.481 | 21 | STN | 15.594 | 91 |
| REP(STN) | 233.083 | 79 | REP(STN) | 1.534 | 9 |

APPENDIX C(2): Metal and nutrient concentrations from three stations (CS126, CS164, CS167) where replicate cores were collected from a single box core to investigate within station variability. Metal and nutrient concentrations from station CS135 where two samples were split into three subsamples to investigate within sample variability.

ND = No data

APPENDIX C(3i): Sediment concentrations (ug/g dry wt.) in QA/QC samples compared with concentrations in routine samples collected from the same stations.
St. Lawrence River, Cornwall, Ontario, 1997. a = 0.05. All data corrected for particle size.

| Station | Al corr | Cd corr | Cr corr | Cu corr | Fe corr | Hg corr | Mn corr | Ni corr | Pb corr | Zn corr | TP corr | TOC corr |
|-------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----------|
| QA/QC CS126-1 | 21783 | 1.012 | 46.8 | 54.8 | 24100 | 17.46 | 420 | 32.6 | 59.8 | 358 | 1003 | 7.361 |
| QA/QC CS126-2 | 17464 | 0.756 | 39.5 | 42.2 | 23088 | 13.42 | 397 | 25.8 | 48.4 | 228 | 1037 | 7.510 |
| QA/QC CS126-3 | 25247 | 1.036 | 70.2 | 53.8 | 29948 | 12.38 | 550 | 44.4 | 55.5 | 319 | 1170 | 7.693 |
| Mean | 21498 | 0.935 | 52.2 | 50.3 | 25712 | 14.42 | 456 | 34.2 | 54.6 | 302 | 1070 | 7.521 |
| Standard Deviation | 3899 | 0.155 | 16.0 | 7.0 | 3703 | 2.68 | 83 | 9.4 | 5.8 | 67 | 88 | 0.166 |
| Cl | 9687 | 0.385 | 39.8 | 17.5 | 9200 | 6.66 | 205 | 23.4 | 14.3 | 166 | 219 | 0.412 |
| 95% Upper Confidence Interval | 31184 | 1.320 | 91.9 | 67.7 | 34912 | 21.08 | 661 | 57.7 | 68.9 | 468 | 1290 | 7.933 |
| 95% Lower Confidence Interval | 11811 | 0.550 | 12.4 | 32.8 | 16513 | 7.76 | 250 | 10.8 | 40.3 | 136 | 851 | 7.109 |
| CS126 | 17381 | 1.300 | 43.0 | 48.0 | 22922 | 9.39 | 363 | 34.5 | 45.5 | 226 | 825 | 5.552 |

| Station | Al corr | Cd corr | Cr corr | Cu corr | Fe corr | Hg corr | Mn corr | Ni corr | Pb corr | Zn corr | TP corr | TOC corr |
|-------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----------|
| QA/QC CS164-1 | 15187 | 1.035 | 40.0 | 58.1 | 15864 | 3.14 | 268 | 25.3 | 180.9 | 566 | 610 | 2.565 |
| QA/QC CS164-2 | 15288 | 0.907 | 42.9 | 60.8 | 15475 | 3.69 | 261 | 26.3 | 175.4 | 598 | 593 | 2.413 |
| QA/QC CS164-3 | 15148 | 1.004 | 42.0 | 55.5 | 15557 | 3.27 | 266 | 29.7 | 122.8 | 583 | 599 | 2.578 |
| Mean | 15208 | 0.982 | 41.6 | 58.1 | 15632 | 3.36 | 265 | 27.1 | 159.7 | 583 | 601 | 2.519 |
| Standard Deviation | 72 | 0.067 | 1.4 | 2.7 | 205 | 0.28 | 4 | 2.3 | 32.1 | 16 | 9 | 0.092 |
| Cl | 179 | 0.166 | 3.6 | 6.6 | 509 | 0.71 | 9 | 5.7 | 79.6 | 40 | 22 | 0.228 |
| 95% Upper Confidence Interval | 15387 | 1.148 | 45.2 | 64.7 | 16141 | 4.07 | 274 | 32.8 | 239.3 | 623 | 623 | 2.747 |
| 95% Lower Confidence Interval | 15028 | 0.816 | 38.1 | 51.5 | 15123 | 2.66 | 256 | 21.4 | 80.1 | 542 | 578 | 2.291 |
| CS164 | 16676 | 0.926 | 50.8 | 62.4 | 18448 | 3.19 | 304 | 26.0 | 141.7 | 697 | 883 | 2.558 |

| Station | Al corr | Cd corr | Cr corr | Cu corr | Fe corr | Hg corr | Mn corr | Ni corr | Pb corr | Zn corr | TP corr | TOC corr |
|-------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----------|
| QA/QC CS167-1 | 16455 | 0.978 | 33.8 | 32.4 | 15985 | 1.03 | 245 | 26.8 | 31.9 | 102 | 722 | 2.663 |
| QA/QC CS167-2 | 17212 | 1.040 | 35.5 | 32.1 | 16823 | 1.19 | 255 | 26.2 | 33.8 | 107 | 654 | 2.766 |
| QA/QC CS167-3 | 18422 | 1.097 | 37.2 | 34.3 | 18112 | 0.87 | 283 | 28.9 | 31.8 | 110 | 635 | 2.934 |
| Mean | 17363 | 1.038 | 35.5 | 32.9 | 16973 | 1.03 | 261 | 27.3 | 32.5 | 106 | 671 | 2.788 |
| Standard Deviation | 992 | 0.060 | 1.7 | 1.2 | 1072 | 0.16 | 19 | 1.4 | 1.2 | 4 | 46 | 0.137 |
| Cl | 2465 | 0.148 | 4.2 | 2.9 | 2662 | 0.39 | 48 | 3.5 | 2.9 | 9 | 113 | 0.340 |
| 95% Upper Confidence Interval | 19828 | 1.187 | 39.7 | 35.8 | 19635 | 1.42 | 309 | 30.8 | 35.4 | 115 | 784 | 3.128 |
| 95% Lower Confidence Interval | 14898 | 0.890 | 31.2 | 30.0 | 14311 | 0.64 | 213 | 23.7 | 29.6 | 97 | 557 | 2.447 |
| CS167 | 17243 | 0.934 | 38.0 | 40.7 | 16730 | 1.22 | 272 | 30.7 | 37.6 | 110 | 890 | 3.048 |

| Station | Al corr | Cd corr | Cr corr | Cu corr | Fe corr | Hg corr | Mn corr | Ni corr | Pb corr | Zn corr | TP corr | TOC corr |
|-------------------------------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|---------|----------|
| QA/QC CS135-1 | 17144 | 0.967 | 37.0 | 40.4 | 16493 | 12.26 | 305 | 28.5 | 40.4 | 306 | 780 | 2.589 |
| QA/QC CS135-2 | 18810 | 0.926 | 37.3 | 41.3 | 18134 | 346 | 29.1 | 43.5 | 297 | 766 | 2.485 | |
| QA/QC CS135-3 | 18109 | 0.988 | 37.4 | 40.5 | 17215 | 315 | 29.5 | 43.1 | 309 | 665 | 2.667 | |
| Mean | 18021 | 0.960 | 37.2 | 40.7 | 17280 | 12.26 | 322 | 29.0 | 42.3 | 304 | 737 | 2.580 |
| Standard Deviation | 837 | 0.032 | 0.2 | 0.5 | 823 | 21 | 0.5 | 1.7 | 6 | 63 | 0.092 | |
| Cl | 2078 | 0.079 | 0.6 | 1.3 | 2043 | 53 | 1.2 | 4.2 | 15 | 156 | 0.227 | |
| 95% Upper Confidence Interval | 20099 | 1.039 | 37.8 | 42.1 | 19324 | 375 | 30.2 | 46.6 | 318 | 893 | 2.808 | |
| 95% Lower Confidence Interval | 15943 | 0.882 | 36.7 | 39.4 | 15237 | 269 | 27.8 | 38.1 | 289 | 581 | 2.353 | |
| CS135 | 21048 | 1.090 | 44.6 | 50.4 | 20494 | 4.15 | 367 | 25.4 | 65.1 | 259 | 731 | 2.335 |

APPENDIX C(3)(b) Sediment concentrations (ug/g dry wt.) in QA/QC samples compared with concentrations in routine samples collected from the same stations.
St. Lawrence River, Cornwall, Ontario, 1997. $\alpha = 0.05$

| Station | Al | Cd | Cr | Cu | Fe | Hg | Mn | Ni | Pb | Zn | TP | TOC | Sand | Silt | Clay |
|-------------------------------|-------|-------|------|------|-------|-------|------|------|-------|-----|-------|-------|------|------|------|
| QA/QC CS126-1 | 14100 | 0.655 | 30.3 | 35.5 | 15600 | 11.30 | 272 | 21.1 | 38.7 | 232 | 649 | 4.765 | 52.1 | 31.9 | 16.0 |
| QA/QC CS126-2 | 11800 | 0.511 | 26.7 | 28.5 | 15600 | 9.07 | 268 | 17.4 | 32.7 | 154 | 701 | 5.074 | 49.6 | 30.0 | 20.0 |
| QA/QC CS126-3 | 14500 | 0.595 | 40.3 | 30.9 | 17200 | 7.11 | 316 | 25.5 | 31.9 | 183 | 672 | 4.418 | 57.6 | 29.5 | 13.0 |
| Mean | 13487 | 0.59 | 32.4 | 31.6 | 16133 | 9.16 | 285 | 21.3 | 34.4 | 190 | 674 | 4.752 | 53.1 | 30.5 | 16.3 |
| Standard Deviation | 1457 | 0.07 | 7.0 | 3.6 | 924 | 2.10 | 27 | 4.1 | 3.7 | 39 | 26 | 0.328 | 4.1 | 1.3 | 3.5 |
| Cl | 3620 | 0.180 | 17.5 | 8.8 | 2295 | 5.21 | 66 | 10.1 | 9.2 | 98 | 65 | 0.815 | 10.2 | 3.1 | 8.7 |
| 95% Upper Confidence Interval | 17086 | 0.767 | 49.9 | 40.5 | 18428 | 14.37 | 351 | 31.4 | 43.7 | 288 | 739 | 5.568 | 63.3 | 33.6 | 25.1 |
| 95% Lower Confidence Interval | 9847 | 0.407 | 14.9 | 22.8 | 13839 | 3.95 | 219 | 11.3 | 25.2 | 92 | 609 | 3.937 | 42.9 | 27.3 | 7.6 |
| CS126 | 9630 | 0.720 | 23.8 | 26.6 | 12700 | 5.20 | 201 | 19.1 | 25.2 | 125 | 457 | 3.076 | 58.3 | 30.5 | 10.5 |
| Station | Al | Cd | Cr | Cu | Fe | Hg | Mn | Ni | Pb | Zn | TP | TOC | Sand | Silt | Clay |
| QA/QC CS164-1 | 15700 | 1.070 | 41.4 | 60.1 | 16400 | 3.25 | 277 | 26.2 | 187.0 | 585 | 631 | 2.652 | 23.5 | 58.8 | 17.7 |
| QA/QC CS164-2 | 16300 | 0.967 | 45.7 | 64.8 | 16500 | 3.93 | 278 | 28.0 | 187.0 | 638 | 632 | 2.573 | 21.0 | 66.1 | 12.8 |
| QA/QC CS164-3 | 14800 | 0.981 | 41.0 | 54.2 | 15200 | 3.19 | 260 | 29.0 | 120.0 | 570 | 585 | 2.519 | 26.4 | 57.3 | 15.0 |
| Mean | 15600 | 1.006 | 42.7 | 59.7 | 16033 | 3.46 | 272 | 27.7 | 164.7 | 598 | 616 | 2.581 | 23.6 | 60.7 | 15.2 |
| Standard Deviation | 755 | 0.056 | 2.6 | 5.3 | 723 | 0.41 | 10 | 1.4 | 38.7 | 36 | 27 | 0.067 | 2.7 | 4.7 | 2.5 |
| Cl | 1875 | 0.139 | 6.5 | 13.2 | 1797 | 1.02 | 25 | 3.5 | 96.1 | 89 | 67 | 0.166 | 6.7 | 11.7 | 6.1 |
| 95% Upper Confidence Interval | 17475 | 1.145 | 49.2 | 72.9 | 17830 | 4.48 | 297 | 31.3 | 260.8 | 686 | 683 | 2.747 | 30.3 | 72.4 | 21.3 |
| 95% Lower Confidence Interval | 13725 | 0.867 | 36.2 | 46.5 | 14236 | 2.44 | 247 | 24.2 | 68.6 | 509 | 549 | 2.415 | 16.9 | 49.0 | 9.1 |
| CS164 | 16000 | 0.888 | 48.7 | 59.9 | 17700 | 3.06 | 292 | 24.9 | 136.0 | 669 | 847 | 2.454 | 29.0 | 53.6 | 17.4 |
| Station | Al | Cd | Cr | Cu | Fe | Hg | Mn | Ni | Pb | Zn | TP | TOC | Sand | Silt | Clay |
| QA/QC CS167-1 | 17500 | 1.040 | 35.9 | 34.5 | 17000 | 1.10 | 261 | 28.5 | 33.9 | 109 | 768 | 2.832 | 21.3 | 63.4 | 15.3 |
| QA/QC CS167-2 | 17700 | 1.070 | 36.5 | 33.0 | 17300 | 1.22 | 262 | 26.9 | 34.8 | 110 | 673 | 2.844 | 23.9 | 61.5 | 14.6 |
| QA/QC CS167-3 | 17800 | 1.060 | 35.9 | 33.1 | 17500 | 0.84 | 273 | 27.9 | 30.7 | 106 | 614 | 2.835 | 28.5 | 56.0 | 15.5 |
| Mean | 17667 | 1.057 | 36.1 | 33.5 | 17267 | 1.05 | 285 | 27.8 | 33.1 | 108 | 685 | 2.837 | 24.6 | 60.3 | 15.1 |
| Standard Deviation | 153 | 0.015 | 0.3 | 0.8 | 252 | 0.19 | 7 | 0.8 | 2.2 | 2 | 78 | 0.006 | 3.6 | 3.8 | 0.5 |
| Cl | 379 | 0.038 | 0.9 | 2.1 | 625 | 0.48 | 17 | 2.0 | 5.4 | 5 | 193 | 0.016 | 9.1 | 9.5 | 1.2 |
| 95% Upper Confidence Interval | 18046 | 1.095 | 37.0 | 35.6 | 17892 | 1.54 | 282 | 29.8 | 38.5 | 114 | 878 | 2.853 | 33.6 | 69.8 | 18.3 |
| 95% Lower Confidence Interval | 17287 | 1.019 | 35.2 | 31.5 | 16642 | 0.57 | 249 | 25.8 | 27.8 | 103 | 492 | 2.821 | 15.5 | 50.8 | 14.0 |
| CS167 | 16800 | 0.910 | 37.0 | 39.7 | 16300 | 1.19 | 265 | 29.9 | 36.6 | 107 | 867 | 2.970 | 27.9 | 51.6 | 20.5 |
| Station | Al | Cd | Cr | Cu | Fe | Hg | Mn | Ni | Pb | Zn | TP | TOC | Sand | Silt | Clay |
| QA/QC CS135-1 | 15800 | 0.891 | 34.1 | 37.2 | 15200 | 11.30 | 281 | 26.3 | 37.2 | 282 | 719 | 2.386 | 31.8 | 49.5 | 18.7 |
| QA/QC CS135-2 | 16700 | 0.822 | 33.1 | 36.7 | 16100 | 30.7 | 25.8 | 38.6 | 264 | 680 | 2.206 | 34.4 | 49.1 | 16.6 | |
| QA/QC CS135-3 | 16200 | 0.884 | 33.5 | 36.2 | 15400 | 282 | 26.4 | 38.6 | 276 | 595 | 2.386 | 33.8 | 51.4 | 14.8 | |
| Mean | 16233 | 0.866 | 33.6 | 36.7 | 15567 | 11.30 | 290 | 26.2 | 38.1 | 274 | 665 | 2.326 | 33.3 | 50.0 | 16.7 |
| Standard Deviation | 451 | 0.038 | 0.5 | 0.5 | 473 | 15 | 0.3 | 0.8 | 9 | 63 | 0.104 | 1.4 | 1.2 | 2.0 | |
| Cl | 1120 | 0.094 | 1.3 | 1.2 | 1174 | 37 | 0.8 | 2.0 | 23 | 158 | 0.258 | 3.4 | 3.1 | 4.8 | |
| 95% Upper Confidence Interval | 17353 | 0.960 | 34.8 | 37.9 | 16741 | 327 | 27.0 | 40.1 | 297 | 822 | 2.584 | 36.7 | 53.1 | 21.5 | |
| 95% Lower Confidence Interval | 15113 | 0.771 | 32.3 | 35.5 | 14393 | 253 | 25.4 | 36.1 | 251 | 507 | 2.068 | 30.0 | 46.9 | 11.9 | |
| CS135 | 19000 | 0.984 | 40.3 | 45.5 | 18500 | 3.75 | 331 | 22.9 | 58.8 | 234 | 660 | 2.108 | 33.2 | 44.9 | 21.9 |

*value from single sample

APPENDIX C(4): QA/QC Sampling. Metal and total phosphorus concentrations (Mg/Kg dry weight), % TOC and particle size distribution in sediment samples (top 10 cm core sample) collected from the St. Lawrence River, 1997

| Station | Replicate core samples | Al Mg/Kg | Cd Mg/Kg | Cr Mg/Kg | Cu Mg/Kg | Fe Mg/Kg | Hg Mg/Kg | Mn Mg/Kg | Ni Mg/Kg | Pb Mg/Kg | Zn Mg/Kg | TP Mg/Kg | TOC % | Sand % | Silt % | Clay % |
|-----------------------------|------------------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|-------------|----------|-----------|-----------|-----------|
| CS126-1 | | 14100 | 0.655 | 30.3 | 35.5 | 15600 | 11.30 | 272 | 21.1 | 38.7 | 232 | 649 | 4.765 | 52.1 | 31.9 | 16.0 |
| CS126-2 | | 11890 | 0.511 | 26.7 | 28.5 | 15600 | 9.07 | 268 | 17.4 | 32.7 | 154 | 701 | 5.074 | 49.6 | 30.0 | 20.0 |
| CS126-3 | | 14500 | 0.595 | 40.3 | 30.9 | 17200 | 7.11 | 316 | 25.5 | 31.9 | 183 | 672 | 4.418 | 57.6 | 29.5 | 13.0 |
| CS164-1 | | 15700 | 1.070 | 41.4 | 60.1 | 16400 | 3.25 | 277 | 26.2 | 187.0 | 585 | 631 | 2.652 | 23.5 | 58.8 | 17.7 |
| CS164-2 | | 16300 | 0.967 | 45.7 | 64.8 | 16500 | 3.93 | 278 | 28.0 | 187.0 | 638 | 632 | 2.573 | 21.0 | 66.1 | 12.8 |
| CS164-3 | | 14800 | 0.981 | 41.0 | 54.2 | 15200 | 3.19 | 260 | 29.0 | 120.0 | 570 | 585 | 2.519 | 26.4 | 57.3 | 15.0 |
| CS167-1 | | 17500 | 1.040 | 35.9 | 34.5 | 17000 | 1.10 | 261 | 28.5 | 33.9 | 109 | 768 | 2.832 | 21.3 | 63.4 | 15.3 |
| CS167-2 | | 17700 | 1.070 | 36.5 | 33.0 | 17300 | 1.22 | 262 | 26.9 | 34.8 | 110 | 673 | 2.844 | 23.9 | 61.5 | 14.6 |
| CS167-3 | | 17800 | 1.060 | 35.9 | 33.1 | 17500 | 0.84 | 273 | 27.9 | 30.7 | 106 | 614 | 2.835 | 28.5 | 56.0 | 15.5 |
| CS135-1 | | 15800 | 0.891 | 34.1 | 37.2 | 15200 | 11.30 | 281 | 26.3 | 37.2 | 282 | 719 | 2.386 | 31.8 | 49.5 | 18.7 |
| <i>Within sample splits</i> | | | | | | | | | | | | | | | | |
| CS135-2 | | 16700 | 0.822 | 33.1 | 36.7 | 16100 | | 307 | 25.8 | 38.6 | 264 | 680 | 2.206 | 34.4 | 49.1 | 16.6 |
| CS135-2 | | 15500 | 0.834 | 31.8 | 35.4 | 14900 | | 280 | 25.6 | 40.8 | 252 | 592 | 2.209 | 35.1 | 47.8 | 17.1 |
| CS135-2 | | 17700 | 0.941 | 36.9 | 42.7 | 16600 | | 308 | 28.6 | 49.5 | 325 | 890 | 2.776 | 28.1 | 53.7 | 18.2 |
| CS135-3 | | 16200 | 0.884 | 33.5 | 36.2 | 15400 | | 282 | 26.4 | 38.6 | 276 | 595 | 2.386 | 33.8 | 51.4 | 14.8 |
| CS135-3 | | 16300 | 0.936 | 33.0 | 39.5 | 15300 | | 278 | 28.3 | 48.5 | 288 | 707 | 2.399 | 35.0 | 45.0 | 20.1 |
| CS135-3 | | 18100 | 1.290 | 35.3 | 40.7 | 16500 | | 318 | 27.5 | 44.1 | 306 | 744 | 2.519 | 28.8 | 50.7 | 20.5 |

APPENDIX D: Component loadings and percent of total variance explained for the PCA of sediment quality in the St. Lawrence River, 1997.

Log Transformed Data - using only percent silt

| | PC I | PC II | PC III |
|------------|--------|--------|--------|
| Aluminum | 0.952 | -0.151 | 0.06 |
| Cadmium | 0.84 | 0.293 | -0.028 |
| Chromium | 0.961 | 0.179 | 0.138 |
| Copper | 0.644 | 0.714 | 0.005 |
| Iron | 0.912 | -0.3 | 0.014 |
| Mercury | -0.601 | 0.658 | -0.216 |
| Manganese | 0.793 | -0.438 | 0.065 |
| Nickel | 0.93 | 0.011 | -0.034 |
| Phosphorus | 0.799 | -0.308 | 0.178 |
| Lead | 0.151 | 0.918 | 0.271 |
| Zinc | 0.21 | 0.923 | 0.145 |
| TOC | 0.486 | 0.175 | -0.831 |
| Silt | 0.872 | 0.115 | -0.163 |

Percent of total variance explained

56% 81% 88%

APPENDIX E: Data from the MOE and EC interlaboratory comparison.

MOE SPLIT SAMPLE DATA

| Station Number (EC Station Description) | MOE Station Description | MOE Field Sample No. | MOE Station Sample No. | Aluminum mg/Kg | Cadmium mg/Kg | Chromium mg/Kg | Copper mg/Kg | Iron mg/Kg | Mercury mg/Kg | Manganese0 mg/Kg | Nickel mg/Kg | Lead mg/Kg | Zinc mg/Kg | SUM1 Clay | SUM2 Silt | SUM3 Sand |
|--|----------------------------|-------------------------|---------------------------|-------------------|------------------|-------------------|-----------------|---------------|------------------|---------------------|-----------------|---------------|---------------|-----------|-----------|-----------|
| CS167-SPLIT | 194 | GL778804 | 12000 | 1.6 | 35 | 41 | 18000 | 0.83 | 250 | 25 | 36 | 130 | 14.0 | 44 | 42 | |
| CS181-SPLIT | 199 | GL778802 | 15000 | 1.4 | 38 | 37 | 22000 | 0.10 | 390 | 28 | 32 | 130 | 17.9 | 40 | 42 | |
| CS168-SPLIT | 195 | GL778803 | 8300 | 1.2 | 28 | 38 | 14000 | 1.50 | 210 | 23 | 36 | 100 | 9.3 | 36 | 55 | |
| CS171-SPLIT | 197 | GL778801 | 10000 | 1.0 | 28 | 30 | 16000 | 0.26 | 300 | 21 | 29 | 120 | 14.5 | 35 | 50 | |
| CS105-SPLIT | 162 | GL778805 | 13000 | 1.4 | 44 | 61 | 20000 | 1.40 | 300 | 28 | 76 | 710 | 20.7 | 58 | 21 | |
| CS164-SPLIT | 192 | GL778806 | 11000 | 1.2 | 46 | 75 | 16000 | 2.40 | 270 | 25 | 170 | 780 | 17.4 | 52 | 30 | |
| CS135-SPLIT | 157 | GL778808 | 11000 | 1.3 | 36 | 57 | 16000 | 5.80 | 300 | 24 | 51 | 400 | 26.3 | 60 | 13 | |
| CS126-SPLIT | 183 | GL778809 | 8900 | 1.0 | 31 | 39 | 16000 | 5.50 | 280 | 22 | 38 | 210 | 15.9 | 52 | 32 | |

EC SPLIT SAMPLE DATA

| Station Number (EC Station Description) | Aluminum mg/Kg | Cadmium mg/Kg | Chromium mg/Kg | Copper mg/Kg | Iron mg/Kg | Mercury mg/Kg | Manganese mg/Kg | Nickel mg/Kg | Lead mg/Kg | Zinc mg/Kg | SUM1 Clay | SUM2 Silt | SUM3 Sand |
|--|-------------------|------------------|-------------------|-----------------|---------------|------------------|--------------------|-----------------|---------------|---------------|-----------|-----------|-----------|
| CS167-SPLIT | 20400 | 1.160 | 42.3 | 37.9 | 19800 | 0.74 | 290 | 30.0 | 38.1 | 125 | 20.1 | 54.3 | 25.7 |
| CS181-SPLIT | 22400 | 0.984 | 41.7 | 33.6 | 21400 | 0.14 | 381 | 30.7 | 29.6 | 112 | 25.0 | 60.9 | 14.0 |
| CS168-SPLIT | 15000 | 0.598 | 36.1 | 33.1 | 15700 | 1.18 | 266 | 24.3 | 33.2 | 101 | 15.7 | 39.4 | 44.9 |
| CS171-SPLIT | 16500 | 0.719 | 33.1 | 29.8 | 17000 | 0.37 | 310 | 23.4 | 25.5 | 109 | 18.3 | 37.7 | 43.9 |
| CS105-SPLIT | 19000 | 1.020 | 43.4 | 45.7 | 18400 | 1.95 | 286 | 26.0 | 69.7 | 581 | 21.6 | 68.6 | 9.8 |
| CS164-SPLIT | 15300 | 0.993 | 44.9 | 55.3 | 15900 | 2.80 | 269 | 25.1 | 135.0 | 620 | 15.7 | 50.3 | 33.2 |
| CS135-SPLIT | 19100 | 0.948 | 40.2 | 45.7 | 17000 | 9.41 | 315 | 26.3 | 46.1 | 345 | 22.3 | 43.3 | 34.4 |
| CS126-SPLIT | 12000 | 0.615 | 26.7 | 37.0 | 14800 | 6.30 | 260 | 18.1 | 29.4 | 142 | 11.0 | 26.0 | 62.6 |

